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Priming of Soil Carbon Decomposition in Two Inner Mongolia Grassland Soils following Sheep Dung Addition: A Study Using ^{13}C Natural Abundance Approach

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Abstract

To investigate the effect of sheep dung on soil carbon (C) sequestration, a 152 days incubation experiment was conducted with soils from two different Inner Mongolian grasslands, *i.e.* a *Leymus chinensis* dominated grassland representing the climax community (2.1% organic matter content) and a heavily degraded *Artemisia frigida* dominated community (1.3% organic matter content). Dung was collected from sheep either fed on *L. chinensis* (C_3 plant with $\delta^{13}\text{C} = -26.8\text{‰}$; dung $\delta^{13}\text{C} = -26.2\text{‰}$) or *Cleistogenes squarrosa* (C_4 plant with $\delta^{13}\text{C} = -14.6\text{‰}$; dung $\delta^{13}\text{C} = -15.7\text{‰}$). Fresh C_3 and C_4 sheep dung was mixed with the two grassland soils and incubated under controlled conditions for analysis of ^{13}C - CO_2 emissions. Soil samples were taken at days 17, 43, 86, 127 and 152 after sheep dung addition to detect the $\delta^{13}\text{C}$ signal in soil and dung components. Analysis revealed that 16.9% and 16.6% of the sheep dung C had decomposed, of which 3.5% and 2.8% was sequestered in the soils of *L. chinensis* and *A. frigida* grasslands, respectively, while the remaining decomposed sheep dung was emitted as CO_2 . The cumulative amounts of C respired from dung treated soils during 152 days were 7–8 times higher than in the un-amended controls. In both grassland soils, ca. 60% of the evolved CO_2 originated from the decomposing sheep dung and 40% from the native soil C. Priming effects of soil C decomposition were observed in both soils, *i.e.* 1.4 g and 1.6 g additional soil C kg^{-1} dry soil had been emitted as CO_2 for the *L. chinensis* and *A. frigida* soils, respectively. Hence, the net C losses from *L. chinensis* and *A. frigida* soils were 0.6 g and 0.9 g C kg^{-1} soil, which was 2.6% and 7.0% of the total C in *L. chinensis* and *A. frigida* grasslands soils, respectively. Our results suggest that grazing of degraded Inner Mongolian pastures may cause a net soil C loss due to the positive priming effect, thereby accelerating soil deterioration.

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Introduction

The availability of soil organic carbon (C) for microbial decomposition is crucial for many processes within the C cycle, and assessment of soil dynamics is of great concern in terms of climate change and soil fertility [1]. Animal dung returned to soil can constitute important source of C, and maintain long-term soil fertility in grassland ecosystems [2–4]. However, dung application can also potentially increase soil respiration [5–8]. Studies have shown that the addition of easily degradable C to soil may stimulate microbial activity to such an extent that the turnover of soil organic matter (SOM) is accelerated temporarily, an effect that is frequently called the priming effect (PE) [3,9–12]. When a positive PE occurs, the addition of material such as animal slurry to soils may not result in a net C sequestration, but rather a net C loss [9]. The intensity, direction, and extent of PE depends on several parameters, including the amount and quality of added C,

soil microbial activity and community structure [13–15], soil pH [16] and aggregate size [17].

Distinct signatures in ^{13}C content between native soil C and ‘new’ introduced labile C compounds added in the form of animal slurry or manure enables quantification of the interaction in C turnover between different C pools [3,18–20]. By this means, researchers have shown that incorporated slurry-C was lost twice as fast as the native soil C in two soils with different C contents. Slurry incorporation induced a PE, which was most pronounced in the soil with the highest C content [18]. Following the application of slurries with different particle sizes to a grassland soil, significant increases of soil CO_2 effluxes (by 2–8 times) were observed in all slurry fractions and the highest was found in the smaller slurry particles [3]. In a study with additions of different substrate quality combinations and C-13 characteristics, Kuzyakov & Bol (2004) have identified three distinct C sources for soil CO_2 emissions and observed that addition of labile C (sugar) lead to changes in SOM

[21]. The ^{13}C natural abundance trace technique has also been applied in the two-phase model of CO_2 emission after dairy or pig slurry application, the first phase (0–48 h) dominated by the incorporation of labile slurry C from the liquid phase, while beyond 48 h slurry-derived C was mainly from less mobile particulate C [12,18,22–23]. However, whereas previous studies mainly focused on the decomposition of cattle dung or slurry [3,5,6,24] or pig slurry [25], less information is available concerning decomposition of sheep faeces C [26].

Inner Mongolia's grasslands in Northern China are representative of large areas of the Eurasian steppe belt [27]. Sheep is the primary livestock in Inner Mongolia's grassland and large amount of sheep dung is applied as fertilizer except for cooking energy [28]. More than 70% of the natural Inner Mongolian grassland area is extensively degenerated as a consequence of increases in livestock numbers and the change of farming systems during the last three decades [28]. Knowledge regarding the importance of sheep dung excretion for soil C cycling in this ecosystem remains sparse. Addition of artificial sheep excreta to Inner Mongolian steppe in autumn did not impact soil microbial biomass C, but microbial activity significantly increased [29–30]. None of these studies, however, have achieved detailed information about the fates of sheep dung derived C in the Inner Mongolian grassland soils, and to what extent heavy grazing by sheep, and thus deposition of dung C, will affect the overall soil C balance through increased sequestration or losses due to priming.

We hypothesized that sheep dung additions to Inner Mongolian grassland soils i) cause a positive priming effect on soil C turnover, and ii) lead to differentiated net C loss in soils with contrasting SOM content. The effects of interaction between soil type and sheep dung addition on soil respiration and soil C sequestration were investigated using the ^{13}C natural abundance technique through a five-month incubation experiment.

The objectives of the study were (1) to assess the input of sheep dung-derived C to the soil C-pool; (2) to examine the extent of SOM priming due to application of dung C; and (3) to quantify the net C pool changes in soils subject to intensive sheep dung application.

Materials and Methods

Ethics Statement

There was no activities involved the endangered or protected species in this study, and Sheep (vertebrates) were involved in this study, so the permission to use sheep in this study was granted from local authority (Dr. Yongfei Bai, Director of Inner Mongolia Grassland Research Station) before we took the soil sample in the field.

Field site details

Soil incubated in the study was taken from the Inner Mongolia Grassland Ecosystem Research Station, Chinese Ecosystem Research Network (IMGERS, 116°42'E, 43°38'N). This region is part of the temperate semiarid steppe belt of Eurasia. The mean annual temperature is slightly above zero (0.8°C) with a January mean of -21°C (absolute minimum -42°C) and a July mean of 19°C (absolute maximum 39°C). Mean annual precipitation is 330.3 mm but fluctuates greatly among years, and most rainfall events occur in July and August (both means for years 1982–2007; IMGERS weather data). The annual frost-free period generally lasts 90–110 days [31].

Field sampling, dung collection and preparation

The soil was collected from two grassland sites. One site was characterized by *Leymus chinensis* vegetation (C_3 plant), which is a dominant species of the climax grassland community in the Inner Mongolian steppe [32]. Under moderate grazing, *L. chinensis* vegetation is replaced by *Cleistogenes squarrosa* (C_4 plant), which in turn is replaced finally by *Artemisia frigida* (C_3 plant) vegetation under heavy grazing [33–34]. About 87% of plants in this geographical region possess a C_3 photosynthetic pathway [33], and consequently the ^{13}C isotopic signature in soil organic carbon reflects a C_3 dominated community (Table 1). Further details on the site conditions are given by Chen & Wang [32].

Soils were collected from 0–10 cm depth in the two grasslands. Five $10 \times 10 \text{ m}^2$ plots situated 2 m apart were sampled randomly using a 6 cm diameter auger to achieve ca. 10 kg of soil from each plot. Soil was composited into one bulk sample, vegetation and coarse roots were removed by hand, and the soil was sieved to pass a 2 mm mesh and stored at $<5^\circ\text{C}$ under field moist conditions until it was used.

Sheep dung was collected from twenty Mongolian sheep (two-year old), housed in metabolic cages with the approval of the Chinese Experimental Animal Committee of the Chinese Academy of Sciences and the owner of sheep. Dung collection did not influence sheep feeding but limited their activities freely in successive ten days by the metabolic cages. The sheep were divided randomly into two groups (ten for each group), and one group was fed on *L. chinensis*, while another group was fed on *C. squarrosa*. After 5 days trial to allow for equilibration of the ^{13}C content in the digestive tract, dung was collected twice per day in plastic bags which were attached to the tails of the sheep. Dung samples were collected during the following five consecutive days. The dung was kept frozen (-20°C) until used. The difference of the C_3 and C_4 sheep dung are given in Table 2. There was no significant different of C, N content and C/N ratios for C_3 and C_4 sheep dung.

Experimental setup

The incubation experiment was conducted over a 5-month period. Before the start of incubation, soil was mixed thoroughly and the soil moisture was adjusted with demineralized water to 40% of water-holding capacity (WHC). The sheep dung was thawed and homogenized by hand before being mixed into the soil. The incubation experiment included six treatments, i.e. the full combination of the two soils and three applications of sheep dung (C_3 , C_4 and no addition). Dung was added as 60 g fresh weight portions mixed thoroughly with 1 kg of soil (air-dried equivalent). The soil-dung mixtures were transferred to 2 l Kilner jars in triplicate that were gently tapped on the lab bench to compress the soil. Two sets of jars were prepared, one for gas sampling and one for soil sampling. To minimize water losses from the soils, jars were covered with perforated Parafilm that was only removed 30 min before gas sampling events. The jars were incubated at $20 \pm 1^\circ\text{C}$ in a controlled temperature cabinet throughout the 152 days of the experiment. Water content was held constant by regular watering to weight.

Headspace sampling for analysis of CO_2 flux and $\delta^{13}\text{C}$ of CO_2

Samples for CO_2 efflux and d^{13}C isotopic analysis were collected 16 times on days 1, 2, 4, 6, 9, 14, 19, 24, 41, 55, 71, 83, 100, 121, 137 and 152 after sheep dung amendment. The Kilner jars were sealed gas tight by lids equipped with a rubber septum to allow headspace gas to be sampled by syringe and

Table 1. Physical and chemical characteristics of the two grassland soils used in the incubation study.

Soil	<i>L. chinensis</i> soil	<i>A. frigida</i> soil
Dominant vegetation	<i>Leymus chinensis</i> , <i>Stipa grandis</i>	<i>Artemisia frigida</i> , <i>Cleistogenes squarrosa</i>
Soil type	Dark chestnut	Light chestnut
Soil texture	Silty loam	Sandy
Soil organic matter (SOM) (%)	2.1±0.06	1.3±0.05
Soil total N content (g kg ⁻¹)	1.9±0.1	1.3±0.1
Soil microbial biomass C (mg/kg ⁻¹)	390±13	251±17
pH (water: soil = 2.5:1)	6.3±0.02	6.6±0.04
δ ¹³ value (‰ vs VPDB)	-22.2±0.1	-22.4±0.8

Numbers are mean ± SE of n = 5 replicate bulk soil samples.
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needle. At each gas sampling event, the headspace was sealed for 30 min and four 10-ml headspace samples were collected every 10 min. The sampling involved a three-step procedure. First, the headspace gas was mixed with a 20-ml sampling syringe several times. Second, a 10-ml gas sample was extracted from the headspace and 5 ml used to pressurize an evacuated 2-ml crimp-sealed vial for the ¹³C isotopic analysis of CO₂. The residual 5 ml gas sample was analyzed immediately for CO₂ concentration by gas chromatography using a HP 6890 GC equipped with a Chromosorb 101 column (30°C), He carrier gas and Thermal Conductivity Detection. Gas fluxes were calculated from the change in CO₂ concentration inside the Kilner jars over the 30 min enclosure period. The relationship between CO₂ concentrations vs. time was significantly linear ($R^2 = 0.93$). Flux rates were thus calculated from the slope of the linear regression lines and expressed as mg C kg⁻¹ soil (DM) day⁻¹.

The ¹³C of CO₂ stored in the 2-ml pressurized vials was determined within 1 week. We used a PreCon (Thermo Scientific, Bremen, Germany) trace gas preparation-concentration unit coupled in continuous flow mode to a Delta PLUS isotope ratio mass spectrometer (IRMS, Thermo Scientific). As laboratory standard we used commercial CO₂ which had been calibrated against certified ¹³CO₂ standards (Messer Griesheim, Krefeld, Germany). Samples of the certified standards were also included in each batch of analysis. Results relating to ¹³C characteristics are reported as ‰ vs. VPDB [35].

Soil sampling and analyses

Soil samples were collected from the jars at days 17, 43, 86, 127 and 152 after dung addition using a 2 cm diameter polyethylene pipe (15 cm long). The subsamples were mixed and the visible small sheep dung was sought out. The larger sheep dung particles

were carefully removed by tweezers and the smaller fractions of sheep residues in the soil was absorbed by electrostatic effect, which produced by a polyethylene bottle rubbing against a piece of fabric [36].

Soil samples were weighed in Ag-foil capsules, arranged on a microtiter plate, wetted with water to approximately field capacity, and placed in a desiccator containing a beaker with concentrated (12M) HCl. The carbonates are released as CO₂ by the acid treatment in 6 to 8 h. The soil samples are then dried at 60°C prior to isotope determination [37]. The soil was then finely ground by a ball mill, and a ca. 30 mg subsample was weighed into a tin combustion cup for determination of total carbon and ¹³C:¹²C ratio following flash combustion on an elemental analyzer (EA 1110, CE Instruments, Milan, Italy) coupled in continuous flow mode to the IRMS.

Calculations

In this study, we assumed that the C₃ and C₄ sheep dung materials go through the same transformation and transport processes, so we can differentiate the carbon source both in soil and CO₂ efflux based on the different isotope value.

We calculated the percentage of dung-derived C in relation to the total soil C according to equation (1):

$$D = \frac{\delta_4 S - \delta_3 S}{\delta_4 d - \delta_3 d} \times 100 \quad (1)$$

where δ₄S and δ₃S are the δ¹³C isotope values of soils amended with C₄ or C₃ dung at the time of sampling, and δ₄d and δ₃d are the δ¹³C isotope values of the original C₄ and C₃ dung prior to amendment [18,38]. The difference in δ¹³C between the C₃ and C₄ dung in our incubation was 10.5‰ (Table 2).

Table 2. Characteristics of sheep dung used in the incubation study. Dung was collected from sheep either fed on *L. chinensis* (C₃ dung) or *C. squarrosa* (C₄ dung).

Dung type	Dry matter	Organic C	Total N	C/N	Dry matter	Total C	δ ¹³ value
	(% w:w)	(% of DM)			(Per 60 g fresh material) ^a		(‰ vs VPDB)
C ₃ dung	86.1±0.4	43.9±0.3 ^b	1.4±0.08	31.3±0.1	51.7±0.4	22.7±0.15	-26.2±0.04
C ₄ dung	85.1±0.6	44.4±0.2	1.29±0.03	34.4±0.1	51.1±0.2	22.7±0.18	-15.7±0.06

^aThe content based on 60g fresh dung portions as applied in the experiment.

^bNumbers are mean ± SE of n = 3 replicates.

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The fractions of dung-derived C incorporated in the soil at the sampling time in relation to the total dung C applied was calculated with equation (2):

$$P = \frac{D \times SDW \times SC \times 100}{DC} \quad (2)$$

where SDW is soil dry weight, SC is the content of soil organic carbon and DC is the amount of dung C added into soil at the beginning of the incubation.

The difference in $\delta^{13}\text{C}$ values between the respired CO_2 from the C_4 and C_3 dung treatments was used to quantify the proportions of dung versus soil-derived CO_2 -C respired from the soil. The dung-derived CO_2 was calculated with equation (3):

$$R = \frac{\delta_4AS - \delta_3AS}{\delta_4d - \delta_3d} \times 100 \times \text{CO}_{2\text{C}4} \quad (3)$$

where $\delta_4AS - \delta_3AS$ is the difference in $\delta^{13}\text{C}$ values of CO_2 emitted from C_4 dung and C_3 dung treatments, and $\text{CO}_{2\text{C}4}$ is the flux of CO_2 in the C_4 dung treatment. The approach assumes that any fractionation in ^{13}C vs. ^{12}C during respiration of sheep dung C and soil C is similar for the C_3 and C_4 dung.

Statistical analyses

Repeated Measures Define Factors of General Linear Model (SPSS 13.0, SPSS Inc. Chicago, Illinois, USA) was used to assess the impacts of treatment, sampling day, soil type and their interactions on the $\delta^{13}\text{C}$ values of soils and CO_2 emission, the CO_2 respired efflux, dung-derived carbon, soil-derived carbon and the primed carbon effect. The sampling day was treated as within-subject variables, and soil type and the treatment were used as a between-subject variable. For each observation of CO_2 emission, cumulative CO_2 emission, $\delta^{13}\text{C}$ of CO_2 emission, dung-derived carbon and soil-derived carbon, the significance of differences between treatments was assessed by two-way ANOVA and Least Significance Difference (LSD).

Results

Soil carbon derived from sheep dung

Independent of soil type, soil $\delta^{13}\text{C}$ did not differ ($p > 0.05$) between the control soil and the C_3 dung amended soil throughout the 152 days period (Table 3). In the C_4 dung soils, $\delta^{13}\text{C}$ significantly exceeded the C_3 and control treatments from day 127 onwards in the *L. chinensis* soil, and from day 43 onwards in the *A. frigida* soil. The temporal incorporation of C_4 sheep dung increased soil $\delta^{13}\text{C}$ values by 0.76‰ from day 17 to day 152 in the *L. chinensis* soil, and by 0.46‰ in the *A. frigida* soil (Table 3).

An increasing amount of dung-derived C appeared in the soil C fractions for both soil types, except for the *L. chinensis* soil at day 86. After 152 days of incubation, about 3.8% and 4.9% of total organic soil C was derived from the applied dung in the *L. chinensis* and *A. frigida* soils, respectively. This was equivalent to 3.5% and 2.8% of the total applied sheep dung C (Table 3).

Daily and cumulative CO_2 fluxes

The addition of sheep dung to the soil significantly increased CO_2 flux throughout the experiment in both soils when compared with the control ($P < 0.05$; Fig. 1 A, B). There was no difference in CO_2 emission between the C_3 dung and C_4 dung treatments for the two soils, except on days 24 and days 100 in the *L. chinensis* soil, where C_3 dung amended soil emitted most CO_2 . The two control soils used in the study showed almost uniform CO_2 emission

patterns, maintaining a constant rate (average $3.2 \text{ mg C kg}^{-1} \text{ soil day}^{-1}$) except for the initial increase in CO_2 flux on days 1–3 which probably resulted from the soil wetting event. A two-phase pattern of soil CO_2 emission was found in the incubation study for both soils. The first phase was observed during 0–55 days after sheep dung amendment, during which CO_2 fluxes decreased to 26% and 43% of the initial CO_2 emission in *L. chinensis* and *A. frigida* soil, respectively. Then a second peak of CO_2 occurred after 55 days, decreasing again after ca. 100 days (*L. chinensis* soil) and 71 days (*A. frigida* soil) (Fig. 1 A, B). There was no interactive effect between soil type and dung treatment during the incubation.

The cumulative CO_2 losses from sheep dung amended soils were ca. 7–8 times higher than in the control soils after 152 days ($P < 0.01$, Fig. 1 C, D). However, there was no difference in total CO_2 emission between the C_3 and C_4 sheep dung treatments ($P > 0.05$, Fig. 1 C, D).

Isotopic characteristics of emitted CO_2

Slightly higher $\delta^{13}\text{C}$ values of CO_2 were found in all the soils included in the two control soils at the beginning of the experiment. These increased towards a peak value at day 6, and then decreased to a minimum asymptotic value in all treatments at day 14 (Fig. 2). Generally, there was no difference in ^{13}C - CO_2 between the C_3 dung treatment and control, except at days 55 and 121 for *A. frigida* soil ($P > 0.05$). However, the $\delta^{13}\text{C}$ value of CO_2 from the C_4 dung treated soils was significantly higher than that in the control soil in most occasions ($P < 0.05$). For the C_4 dung, C_3 dung, and control treatments, respectively, average $\delta^{13}\text{C}$ of CO_2 emissions during 152 days incubation were -14.6‰ , -20.7‰ , and -24.3‰ for *L. chinensis* soil, and -15.2‰ , -20.5‰ , and -24.0‰ for *A. frigida* soil (Fig. 2).

CO_2 emission sources and primed CO_2 emission in dung amended soil

The simple method of estimating the contribution of dung-derived C in respired CO_2 is only valid when the respiration rates from the C_3 and C_4 dung treated soils are the same [38], as was the case in the current study. Two peaks of dung-derived C were observed in both soils (Fig. 3). As a proportion of the total CO_2 -C respired from the dung treated *L. chinensis* soil, during the first 24 days of the experiment, the dung-derived C increased from 11.5% (day 2) to 90.9% (day 24), and from the dung treated *A. frigida* soil increased from 17.9% (day 2) to 91.0% (day 24) (Fig. 3). The cumulative amount of CO_2 -C respired from the C_4 dung treatment was 5.11 and 5.33 g C kg^{-1} for *L. chinensis* soil and *A. frigida* soil, respectively (Appendix S1). The amount of dung-derived C recovered in respiration was nearly 1.5 times that from soil-derived C in both soils, i.e. 59.7% and 58.9% of the evolved CO_2 originated from the decomposing dung in the *L. chinensis* soil and *A. frigida* soil, respectively (Fig. 3).

More CO_2 emissions in C_4 dung treated soil than control soil in this study was ascribed to the priming process. The occurrence of positive priming was observed in both soils, with a more pronounced priming effect in *A. frigida* soil than in *L. chinensis* soil (Fig. 4). Thus, compared with the control soils, an additional 1.34 g C (*L. chinensis*) and 1.55 g C (*A. frigida*) was emitted as CO_2 after sheep dung was applied (Appendix S1 and Table 4).

Net carbon budget

Over the 152 days incubation period, 3.5% and 2.8% (i.e. 0.79 g C and 0.64 g C) of the amended sheep dung was recovered in the *L. chinensis* soil and *A. frigida* soil carbon fractions, respectively (Table 3 and Table 4), and 13.4% and 13.8% of the

Table 3. The dynamics of $\delta^{13}\text{C}$ (Mean \pm SE) in control soils and in soil treated with C_3 and C_4 dung, and the percent of dung-derived C incorporated in the soil in relation to soil C (D) and applied dung C (P).

Days after addition	Control	$\delta^{13}\text{C}$ (‰ vs VPDB)	C_3 dung soil	C_4 dung soil	D (% of soil C)	P (% of applied dung C)
<i>L. chinensis</i> dominated soil						
17	-22.3 ± 0.1	-22.6 ± 0.1	-22.5 ± 0.02	-22.5 ± 0.02	1.4 ± 0.1	1.3 ± 0.1
43	-22.2 ± 0.1	-22.2 ± 0.1	-22.0 ± 0.1	-22.0 ± 0.1	1.1 ± 0.2	1.2 ± 0.1
86	-22.1 ± 0.01	-22.0 ± 0.04	-22.0 ± 0.1	-22.0 ± 0.1	0.5 ± 0.1	0.5 ± 0.01
127	-22.0 ± 0.1^a	-22.1 ± 0.1^a	-21.8 ± 0.1^b	-21.8 ± 0.1^b	3.4 ± 0.2	3.2 ± 0.2
152	-22.2 ± 0.1^a	-22.1 ± 0.3^a	-21.7 ± 0.1^b	-21.7 ± 0.1^b	3.8 ± 0.4	3.5 ± 0.2
<i>A. frigida</i> dominated soil						
17	-22.3 ± 0.03	-22.2 ± 0.1	-22.2 ± 0.03	-22.2 ± 0.03	0.03 ± 0.1	0.02 ± 0.00
43	-22.5 ± 0.1	-22.4 ± 0.1	-22.3 ± 0.04	-22.3 ± 0.04	1.0 ± 0.1	0.6 ± 0.01
86	-22.5 ± 1.5^a	-22.3 ± 0.1^a	-22.1 ± 0.1^b	-22.1 ± 0.1^b	2.1 ± 0.1	1.2 ± 0.04
127	-22.4 ± 0.1^a	-22.3 ± 0.1^a	-21.9 ± 0.2^b	-21.9 ± 0.2^b	4.0 ± 0.3	2.3 ± 0.1
152	-22.2 ± 0.03^a	-22.3 ± 0.01^a	-21.8 ± 0.02^b	-21.8 ± 0.02^b	4.9 ± 0.3	2.8 ± 0.2

Data are shown for each sampling during the 152 days incubation.

Different superscript letters represent statistical significance at $P < 0.05$ at the same sampling time among treatments.

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amended sheep dung was emitted as CO_2 in the *L. chinensis* soil and *A. frigida* soil carbon fractions, respectively.

A priming effect was found in both soils during the incubation period. Considering the apparent supply of dung C to the soil C component, in comparison with the primed CO_2 loss from dung treated soil, the net soil C loss was nearly two times higher from the low C *A. frigida* soil (0.91 g C kg^{-1} soil) compared to the high C

L. chinensis soil (0.55 g C kg^{-1} soil). As a result, 2.6% and 7.0% of soil C was lost due to the application of sheep dung from the *L. chinensis* and *A. frigida* soils, respectively (Table 4).

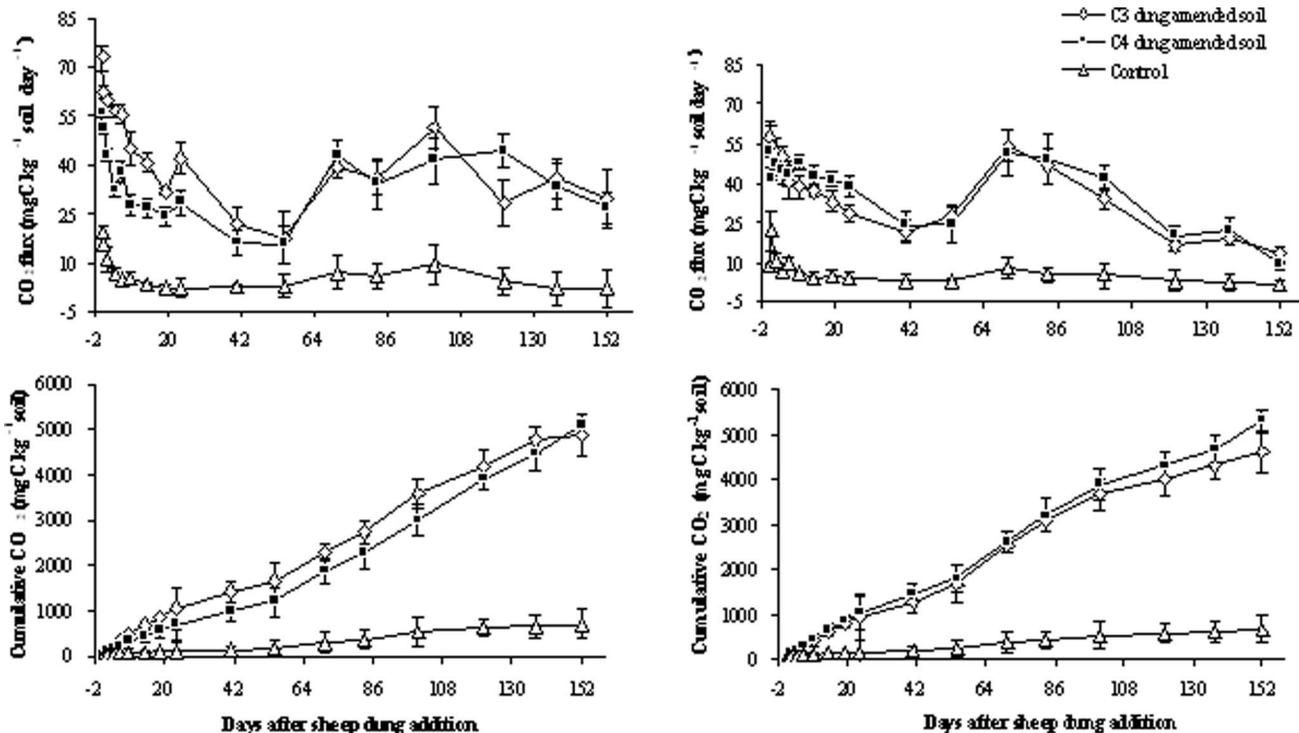


Figure 1. Temporal dynamics of CO_2 fluxes ($\text{mg C kg}^{-1} \text{ day}^{-1}$) (A and B), and cumulative CO_2 -C loss ($\text{mg C kg}^{-1} \text{ day}^{-1}$) (C and D) during 152 days of incubation of *L. chinensis* and *A. frigida* soils amended with C_3 and C_4 dung. Values are the mean ($n=3$) \pm SE (bars).

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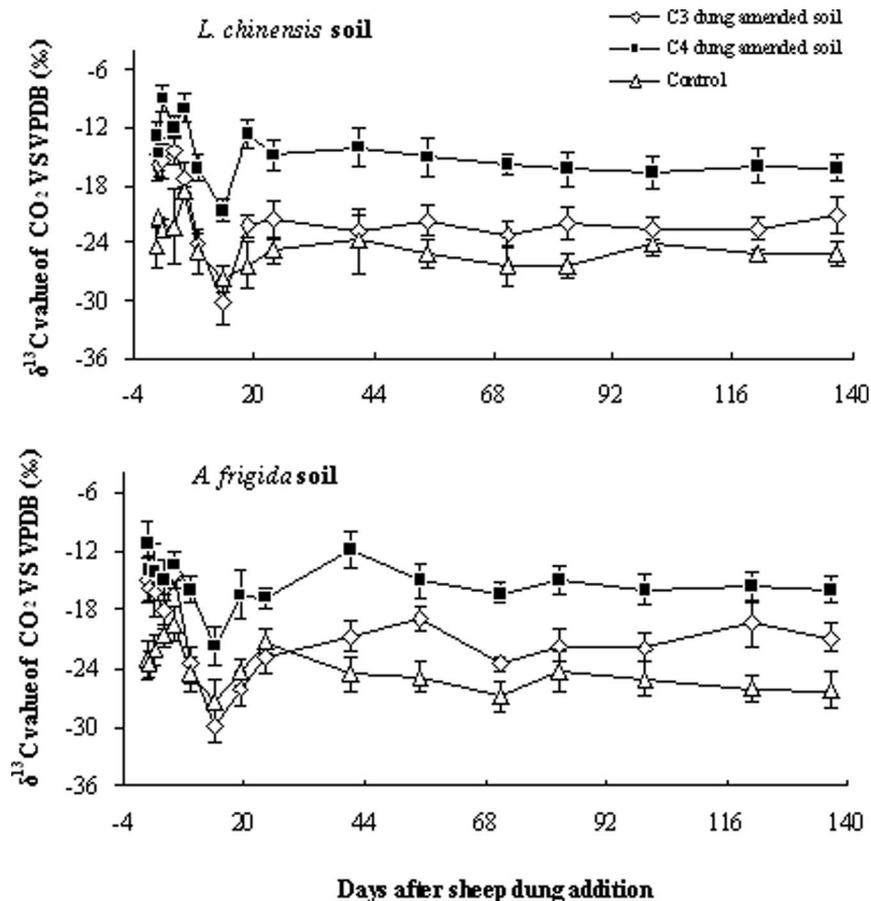


Figure 2. Temporal dynamics of $\delta^{13}\text{C}$ (% vs VPDB) signatures in CO_2 emitted during 152 days following sheep dung addition to *L. chinensis* and *A. frigida* soil ($n=3$).

doi:10.1371/journal.pone.0078578.g002

Discussion

Dung-derived C in the soil

The $\delta^{13}\text{C}$ signatures differed by 10.5‰ between the C_4 and C_3 sheep dung in our study, which makes it possible to differentiate the soil- and dung-derived C in bulk samples as well as in respired CO_2 based on the ^{13}C natural abundance characteristics. Distinct differences in $\delta^{13}\text{C}$ signatures between C_4 and C_3 dung treated (or control) soils emerged 43 days after dung incorporation into *L. chinensis* dominated soil, and 86 days in *A. frigida* dominated soil, suggesting differentiated time lags in the apparent transformation of dung C to the soil C component. However, the limited temporal resolution of soil sampling impeded a detailed identification of the exact temporal dynamics. In contrast, low but significant CO_2 emissions derived from dung C was observed initially in the incubations, which indicates that dung decomposition commenced immediately, but a transfer to the soil C component was not apparent until after several weeks of incubation.

Bol et al. [38] observed that after a 150 days field experiment, 12.6% of applied cattle dung C was retained in a grassland top soil C component. Another field experiment in a temperate grassland showed that a maximum of 60% of cow dung C was retained in the soil after 56 days, declining to around 20% after 372 days [39]. In our study, only 2.8–3.5% of sheep dung C appeared in the soil C component after 152 days of incubation at 20°C. Probably the relatively low dung water content (*i.e.* 14.5%) in our study caused constrained decomposition compared with other studies

(*e.g.* 84% water content in the study by Bol et al. [38]). Slurry generally decomposes faster than dung due to its liquid nature and missing of various dissolved compounds [40]. The characteristics of the dung, such as its C/N ratio, is also an important factor which affected the decomposition of excreta. The C/N ratios in our experiment (31.3 for C_3 dung and 34.4 for C_4 dung) were much higher than the 0.7–10.9 for Bertora et al. [2], and high C/N excreta might be prone to slow mineralization compared to low C/N excreta [2,41–42].

CO_2 fluxes

An apparent two-phase pattern of CO_2 emission was observed in the current study, which was attributed to the two-phase pattern of sheep dung decomposed in both soil.

During the whole period of the incubation (152 days), in both soils, ca 40% of the total CO_2 was released from dung treated soil itself, while 60% released from the decomposed dung (Fig. 3). The two-stage decomposition patterns have also been observed in other studies on dung decomposition, and it is proposed that in the first stage CO_2 emission is due to the decomposition of labile C from soil and easily degradable dung fractions, while in the second phase, the decomposers attack more recalcitrant material [3, 18, 37]. The first phase of decomposition, as indicated by the increased CO_2 efflux, lasted for ca. 6 days, which is longer compared with the 24–48 h duration observed in other studies, suggesting that the labile fraction of sheep dung C is more

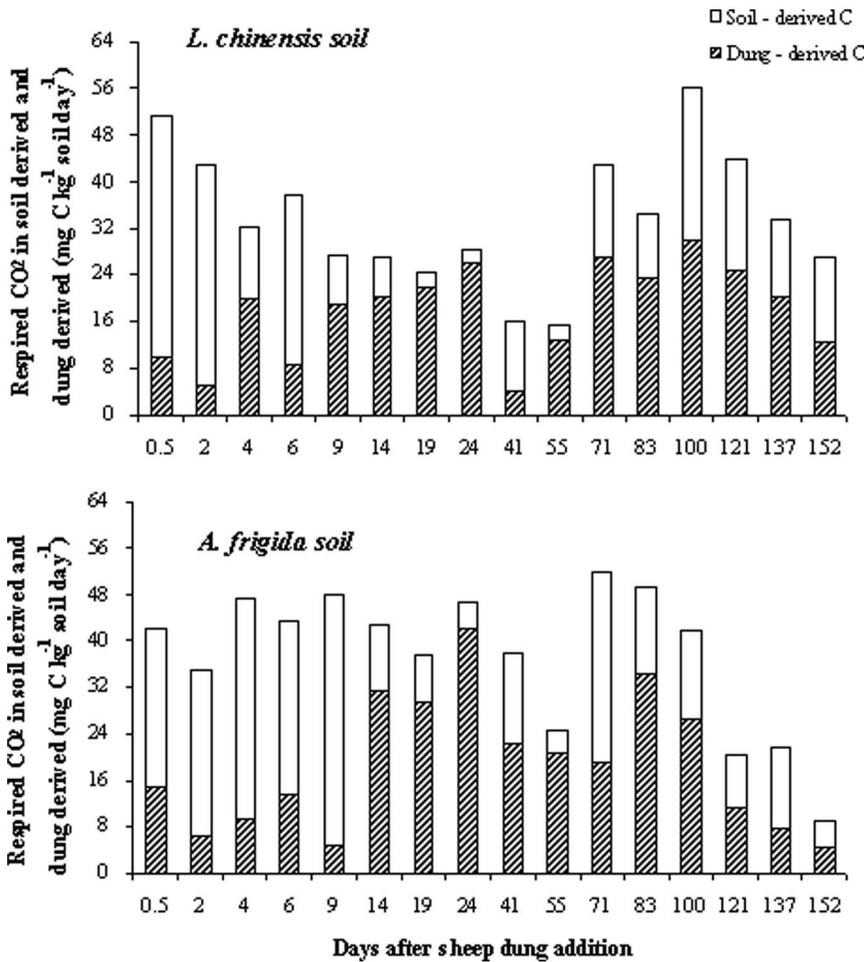


Figure 3. The relative contribution of dung-derived CO₂-C and soil-derived CO₂-C during 152 days incubation calculated from the δ¹³C signature of CO₂ after sheep dung addition (n = 3).
doi:10.1371/journal.pone.0078578.g003

recalcitrant and less available compared to other excreta such as wet cattle dung [24,29].

Isotope characteristics of emitted CO₂

The δ¹³C signal of emitted CO₂ from both C₄ incorporated soils were significantly higher than the C₃ and the control soil in our experiment, which was likely due to the incorporation and the microbial turnover of polymeric biologic cell wall materials from C₄ dung into C₃ grassland soils of C₄ dung C [3]. Similar phenomena have been observed in other studies within the first

hours [25–26]. For the control soil, in general the δ¹³C of emitted CO₂ is slightly higher than the δ¹³C of soil undergoing decomposition. For example, Fangueiro et al. [3] reported that the δ¹³C of CO₂ emitted from untreated soil was on average 5.4‰ higher than the value for the SOM undergoing decomposition. Angers et al. [25] reported 5.0‰ higher δ¹³C of emitted CO₂ than δ¹³C in the soil itself. At the same time, slightly higher δ¹³C of emitted CO₂ in the C₄ dung treated soil (average −14.5‰) than the sheep dung itself (−15.7‰) was found in our incubation, the interaction effect of sheep dung and soil and the isotope

Table 4. Fates of soil and sheep dung carbon after 152 days incubation.

Soil type	Organic C of soil	Fate of sheep dung C		Soil CO ₂ emission (g C kg dry soil ⁻¹)			Net soil C loss	Total soil C loss
		Recovered in soil	Emitted as CO ₂	Soil derived C	Control	Priming		
<i>L. chinensis</i>	21.0±0.06	0.79±0.01	3.05±0.04	2.06±0.01	0.72±0.02	1.34±0.03	0.55±0.01 ^a	2.6±0.2 ^b
<i>A. frigida</i>	13.0±0.05	0.64±0.01	3.14±0.03	2.19±0.01	0.64±0.01	1.55±0.01	0.91±0.02	7.0±0.3

^aNet soil C loss was given by the value of sheep dung C sequestered in the soil during 152 days subtracted from the primed soil CO₂.

^bSoil C loss (%) is the percentage total CO₂-C loss compared to soil total organic C content.

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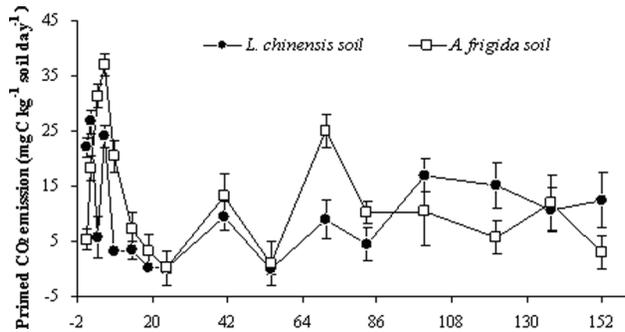


Figure 4. Primed CO₂ emission (mg C kg⁻¹ day⁻¹) during 152 days of incubation from sheep dung amended *L. chinensis* and *A. frigida* soils. The numbers indicate soil C derived CO₂ emitted in excess to the control soil (n=3).

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fractionation associated with the microbial turnover maybe was the possible reason.

Soil priming effects after sheep dung addition

A priming effect (PE) is defined as a short-term change in the turnover of soil organic matter caused by the addition of labile organic C to the soil [43]. Here, we determined the priming effect as the excess emissions of soil C derived CO₂ in dung treated soil compared to control soils. Primed CO₂ emissions were observed throughout the 152 days incubation in both grassland soils. Specifically, 1.3 g and 1.6 g of excess soil C kg⁻¹ was emitted as CO₂ when dung was added to the *L. chinensis* soil and *A. frigida* soil, respectively, corresponding to 6.2% and 11.9% of total soil C. Such a priming effect is in the upper range of primed C losses of 2.3%–8.9% observed in previous studies [44–45].

Priming of soil C decomposition is believed to occur during relatively short-term periods upon addition of labile substrates to soils, and always occurs only in the early stage of substrate addition after which it rapidly ceases [9,10,18,46]. The rapid decrease in the priming effect was likely caused by the depletion of easily decomposable substances added with more complex materials [23]. However, primed soil C decomposition was apparent in our study throughout the entire 152 days period, although with a quantitative variation during the experiment, which may be related to the high C/N ratio and slow decomposition rate characteristics of sheep dung. The priming effect depends not only on the decomposability of the various carbon pools in the environment, but also on the state of the microorganisms [47–48], and involves not only one mechanism but rather a succession of processes partly connected with succession of microbial communities and functions [13]. Further research is needed to test the fundamental processes and mechanisms involved in the priming effect of soil organic matter decomposition due to grazing in Inner Mongolian grasslands.

The loss of sheep dung C via CO₂ was the same in the two grassland soils in our experiment, which contrasts with results by Bol et al. [18] who reported that more slurry-derived C was respired from a C-rich soil compared to a C-poor soil during 0–9 days after slurry incorporation. The authors of that study

attributed this to the more pronounced enhancement of basal soil respiration in C-rich soil compared to the C-poor soil. Furthermore, after labile organic C addition, energy limitation in C-poor soil may have ceased, which subsequently facilitated more activation of soil microorganisms, and more enzymes were produced that were capable of SOM degradation [17,43].

Conclusion and implications for grassland management

The addition of C₄ sheep dung to a C₃ grassland soil enabled us to successfully trace the fate of dung-derived C in the soil and calculate the soil organic C budget for two soils from the Inner Mongolian steppe. Although sheep dung provided an additional organic carbon source for the grassland soils, a large part was emitted as CO₂ to the atmosphere. After sheep dung addition, a positive priming effect of soil C decomposition was observed in both a high-C *L. chinensis* soil and a low-C *A. frigida* soil. Therefore, the balance of soil organic carbon storage was negative when sheep dung was mixed into the soil. This effect was more pronounced for the degraded community of *A. frigida* soils, from which more soil C was lost compared with the climax community of *L. chinensis* soils. This finding is contrary to the conventional conception of carbon storage in temperate grassland, which predicts that the livestock excreta applied to grassland soils could return essential nutrients for plant growth and increase fertility and SOM contents.

The results suggest that intensive grazing management in the temperate steppe should be avoided. In Inner Mongolian grasslands, the succession from *L. chinensis* dominated communities to *A. frigida* dominated communities will result in decreased plant productivity and soil carbon inputs because of the decreased litter and root nutrient return [34]. Bearing in mind that complex plant-soil interactions which exist under field conditions have not been considered in our study, and assuming that the present conclusion can be extrapolated to field conditions, a further acceleration of the decreasing soil C pool in degraded grasslands may occur. Under actual grazing conditions, sheep dung may be mixed merely within the very top grassland soils, suggesting that the current calculations based on well-mixed soil and dung may overestimate soil C losses. Future work is thus needed to examine whether the priming effect of sheep dung amendments observed in this study can be extended to field conditions.

Supporting Information

Appendix S1
(DOC)

Acknowledgments

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Author Contributions

Conceived and designed the experiments: XZM SPW YFW. Performed the experiments: XZM PA. Analyzed the data: XZM. Contributed reagents/materials/analysis tools: XZM CJW. Wrote the paper: XZM.

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