A brief multi-disciplinary review on antimicrobial resistance in medicine and its linkage to the global environmental microbiota.

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The discovery and introduction of antimicrobial agents to clinical medicine was one of the greatest medical triumphs of the 20th century that revolutionized the treatment of bacterial infections. However, the gradual emergence of populations of antimicrobial-resistant pathogenic bacteria resulting from use, misuse, and abuse of antimicrobials has today become a major global health concern. Antimicrobial resistance (AMR) genes have been suggested to originate from environmental bacteria, as clinically relevant resistance genes have been detected on the chromosome of environmental bacteria. As only a few new antimicrobials have been developed in the last decade, the further evolution of resistance poses a serious threat to public health. Urgent measures are required not only to minimize the use of antimicrobials for prophylactic and therapeutic purposes but also to look for alternative strategies for the control of bacterial infections. This review examines the global picture of antimicrobial resistance, factors that favor its spread, strategies, and limitations for its control and the need for continuous training of all stakeholders i.e., medical, veterinary, public health, and other relevant professionals as well as human consumers, in the appropriate use of antimicrobial drugs.

Keywords: antimicrobial resistance, human and veterinary medicine, environment, soil, wastewater, resistance genes
(Levy, 1998), the importance of AMR was already recognized in 1969 both in humans and veterinary medicine as stated in the Swan Report (Swann, 1969). Further research showed that the origin and spread of AMR is, in fact, a very complex problem. Hence, there cannot be a single solution for minimizing AMR; rather a coordinated multi-disciplinary approach will be required to address this issue (Serrano, 2005; Smith et al., 2009). We must also recognize that wherever antimicrobials are used, AMR will inevitably follow.

The purpose of this review is to highlight the problem of resistance to antimicrobials with its consequences, including how the spread of AMR could be limited. We highlight how the numerous useful applications of antimicrobials led to AMR in different ecological locations (Figure 1), aiming to unify the many important aspects of this problem. Finally we advocate the need for teaching and continuous training of all stake-holders (i.e., medical, veterinary, public health, and other relevant professionals) as well as human consumers of antimicrobial drugs, in the appropriate use of antimicrobials.

THE HUMAN MEDICINE AND ANTIMICROBIAL RESISTANCE EMERGENCE OF ANTIMICROBIAL RESISTANCE AND ITS COST

In human medicine AMR is as old as the clinical usage of antimicrobial compounds. Antimicrobial-resistant pathogens have been observed soon after the introduction of new drugs in hospitals where antimicrobials are intensively used (Levy, 1998).
Consequently, AMR in the context of human medicine has dominated the literature for a long time (Figure 2). Over the years, and continuing into the present, almost every known bacterial pathogen and numerous human commensals have developed resistance to one or more antimicrobials in clinical use (Table 1). Extended-spectrum β-lactamases (ESBL) are the ones most often encountered in the hospital (intensive care) setting. Methicillin-resistant Staphylococcus aureus (MRSA) and vancomycin-resistant enterococci (VRE) have also been found to have a significant nosocomial ecology (Otter and French, 2010). In addition, ESBL positive bacteria and MRSA infections are increasingly detected in the community. Furthermore, the increase in fluoroquinolone resistance due to target-site mutations and the worldwide emergence of plasmid-mediated quinolone resistance genes may represent a major challenge in future given the critical importance of this antimicrobial therapy (Cattoir et al., 2007; Strahilevitz et al., 2009). Carbapenems are the last line of defense against the non-Enterobacteriaceae pathogens, such as Pseudomonas aeruginosa and Acinetobacter baumannii (Brown et al., 1998). However, since the first description of the blaOXA genes, there has been a worldwide increase in the dissemination of new resistance determinants conferring carbapenem resistance. For example, the Klebsiella pneumoniae carbapenemase (KPC) type enzymes, Verona integron-encoded metallo-β-lactamase (VIM), Imipenemase Metallo-β-lactamase (IMP) and New Delhi metallo-β-lactamase (NDM), and the OXA-48 type of enzymes have been isolated from a number of bacterial genera irrespective of their geographical distribution (Kumarasamy et al., 2010; Walsh et al., 2011). Carbapenemase resistance mechanisms are found among Escherichia coli and Klebsiella isolates in hospital settings, and to a lesser extent also in the community, thus healthy human carriers begin to be a concern (Nordmann et al., 2012). Furthermore, carbapenemase-producing organisms have also been isolated from farm animals (Fischer et al., 2012).
Table 1 | Antimicrobial resistance detection in some important pathogens soon after arrival of the “magic bullets” into the market.

<table>
<thead>
<tr>
<th>Year</th>
<th>Bacteria</th>
<th>Drug resistance</th>
<th>Comments</th>
<th>References</th>
</tr>
</thead>
<tbody>
<tr>
<td>1948</td>
<td><em>Staphylococcus aureus</em></td>
<td>Penicillin</td>
<td>In British civilian hospitals soon after the introduction of penicillin</td>
<td>Barber and Rozwadowska-Dowzenko, 1948</td>
</tr>
<tr>
<td>1948</td>
<td><em>Mycobacterium tuberculosis</em></td>
<td>Streptomycin</td>
<td>In the community soon after the clinical usage of this antimicrobial</td>
<td>Crofton and Mitchison, 1948</td>
</tr>
<tr>
<td>1950’s–1960’s</td>
<td><em>Escherichia coli</em>, <em>Shigella spp.</em>, and <em>Salmonella enterica</em></td>
<td>Multiple drugs</td>
<td></td>
<td>Watanabe, 1963; Olarte, 1983; Levy, 2001</td>
</tr>
<tr>
<td>1960’s</td>
<td><em>VRE- Enterococcus spp.</em></td>
<td>Multiple drugs</td>
<td></td>
<td>Levy and Marshall, 2004; Nordmann et al., 2011</td>
</tr>
</tbody>
</table>

VRE, Vancomycin-Resistant Enterococcus; ESBL, Extended-spectrum β-lactamase; MRSA, methicillin/oxacillin-resistant *Staphylococcus aureus*; Qβ, Quinolone resistant; MDR, Multi-drug resistant.

HUMAN MOBILITY—THE DIRECT AND INDIRECT IMPACT ON HUMAN PATHOGENS

The increasing cross-border and cross-continental movements of people has a major impact on the spread of multi-resistant bacteria (Linton et al., 1972; Arya and Agarwal, 2011; Cheng et al., 2012). The emergence and global spread of the international clone 1 of penicillin-resistant *Streptococcus pneumoniae* (Klugman, 2002) and the recently occurring New Delhi Metallo-β-lactamase (blaNDM-1) producing *Enterobacteriaceae*, which inactivates all β-lactam antibiotics, including carbapenems, are good examples. The blaNDM-1 appears to have originated in the Indian subcontinent and subsequently could be found in North America, the United Kingdom, and other European countries by the movement of people (Arya and Agarwal, 2011; Walsh et al., 2011).

The AMR problem remains a growing public health concern because infections caused by resistant bacteria are increasingly difficult and expensive to treat. The consequences of this problem are: longer hospital stay, longer time off work, reduced quality of life, greater likelihood of death due to inadequate or delayed treatment, increases in private insurance coverage and an additional cost for hospitals when hospital-acquired infections occur in addition to the increased overall healthcare expenditure (Roberts et al., 2009; Filice et al., 2010; Korczak and Schöffmann, 2010; Wilke, 2010). Thus, in order to calculate the full economic burden of AMR we have to consider the burden of not having antimicrobial treatment options at all, which in the extreme case would probably cause a breakdown of the entire modern medical system (Alanis, 2005; Falagas and Bliziotis, 2007; Pratt, 2010). In short, everyone will be at risk when antimicrobials become ineffective and the threat is greatest for young children, the elderly, and immune-compromised individuals, such as cancer patients undergoing chemotherapy and organ transplant patients (Tablan et al., 2004).

THE VETERINARY MEDICINE AND AGRICULTURE SECTOR CONSUMPTION AND REGULATION OF ANTIMICROBIAL USE IN ANIMALS

The antimicrobials, used in veterinary medicine were introduced soon after they became available for the treatment of human diseases from the mid-1940’s (Gustafson and Bowen, 1997; McEwen, 2006). Even though some drugs are exclusively designed for veterinary use, most belong to the same antimicrobial classes as those used in human medicine with identical or very similar structures (Swann, 1969; Heuer et al., 2009).

Annually, large quantities of drugs are administered to animals in the agricultural sector worldwide to secure a sufficient amount of food (meat, eggs, and dairy products) to feed a rapidly growing world human population (Vazquez-Moreno et al., 1990; Roura et al., 1992; Rassow and Schaper, 1996). As data collection on antimicrobial use in animals was not harmonized to provide reliable and comparable information, and following a request from the European Commission the European Medicines Agency (EMA), the European Surveillance of Antimicrobial Consumption (ESVAC) programme has been created. The ESVAC programme is responsible for collecting, analyzing, and reporting sales data from European countries and developing an organized approach for the collection and reporting of data on antimicrobial use for animals including annual reporting from EU member states (www.emaa.europa.eu/ema/index.jsp?curl=pages/regulation/document_listing/document_listing_000302.jsp). During 2007, in 10 European countries the sale of antimicrobial drugs for therapeutic use as veterinary medicine varied from 18 to 188 mg/kg biomass (Grave et al., 2010).

The administration of antimicrobials to food producing animals can have other purposes than treatment, such as: growth promotion (although now totally banned in Europe and quinolones for the poultry industry are banned in the USA), prophylaxis, and -metaphylaxis (Anthony et al., 2001; Anderson et al., 2003; Casewell et al., 2003; Cabello, 2006). Approximately 70% of all the antimicrobials administered in animal farming are used for non-therapeutic purposes (Roe and Pillai, 2003).

In the European Union (EU) the use of avoparcin was banned in 1997. Furthermore spiramycin, tylosin, and virginiamycin for growth promotion were banned from use in 1998. All other growth promoters in the feed of food-producing animals were banned in the EU—countries from January 1, 2006 (http://europa.eu). Data from Denmark showed that animals could be produced at a large scale without the use of growth promoters,
without adversely affecting the production (Aarestrup et al., 2001; Aarestrup, 2005; Hammerum et al., 2007). In the United States of America, politicians are discussing the introduction of a similar ban on the use of antimicrobials in animal husbandry for growth promotion (http://www.govtrack.us/congress/bills/109/s742). Despite these bans, in some parts of the world, medically important antibiotics are still routinely fed to livestock prophylactically to increase profits and to ward-off potential bacterial infections in the stressed and crowded livestock and aquaculture environments (Cabello, 2006; Smith et al., 2009; Ndii and Barton, 2012). Because stress lowers the function of the immune system in animals, antimicrobials are seen as especially useful in intensive confinements of animals (Halverson, 2000). The non-therapeutic use of antimicrobials involves low-level exposure through feed over long periods—an optimal way in which to enrich resistant bacterial populations (Sharma et al., 2008; Kohanski et al., 2010; Alexander et al., 2011; Gullberg et al., 2011).

Various monitoring programs around the world have started monitoring AMR and a range of research activities and interventions have shown that antimicrobial usage has a large effect upon selection of AMR in animal production. A rapid and parallel decrease in resistant Enterococcus faecium from pig and poultry has been reported in Denmark after the ban of growth promoters in livestock (Aarestrup et al., 2001). The Norwegian aquaculture industry has produced over one million tons of farmed fish (http://www.ssb.no/fiskeoppdrett_en/) by using only 649 kg of antimicrobials in 2011 (NORM/NORM-VET, 2011). It is evident from the Danish integrated AMR monitoring and research program (DANMAP) and NORM/NORM-VET Report (NORM/NORM-VET, 2011) that reduction of antimicrobial usage with strict policies may still be the safest way to control the development and spread of AMR in this sector in the future.

ANTIMICROBIAL-RESISTANT BACTERIA IN COMPANION ANIMALS AND ANIMAL HUSBANDRY

The use of antimicrobials in animal husbandry has for many years actively selected for bacteria which possess genes capable of conferring AMR (Bastianello et al., 1995; Sundin et al., 1995). Consequently, this aspect has also seen much attention in the literature (Figure 2). Despite large differences in methodology, the results of most relevant scientific studies demonstrate that not long after the introduction of antimicrobials in veterinary practice, resistance in pathogenic bacteria, and/or the fecal flora was observed (Caprioli et al., 2000; Jean-Louis et al., 2000). In particular an increased emergence of pathogenic bacteria resistant to antimicrobials has occurred in members of the genera Salmonella, Campylobacter, Listeria, Staphylococcus, Enterococcus, and Escherichia coli. Some resistant strains of these genera are propagated primarily among animals but can subsequently infect people as zoonotic agents (Levy, 1984; Corpet, 1988; Marshall et al., 1990; Giguère et al., 2007).

In veterinary medicine the use of antimicrobials in companion animals such as pets and horses is restricted to therapeutic purposes only. Companion animals are increasingly treated as family members, in the context of applying advanced antimicrobial treatments to their infectious diseases. For instance, skin infections caused by staphylococci in dogs with or without underlying allergic reactions result in an increasing use of semi-synthetic penicillins because of the ineffectiveness of penicillin against penicillinase producing Staphylococcus pseudintermedius (Yoon et al., 2010). Moreover, emerging methicillin-resistant Staphylococcus pseudintermedius (MRSP), methicillin-resistant Staphylococcus aureus (MRSA), and ESBL producing E. coli displaying multidrug resistance has led to increased concern related to AMR in companion animal practice (Bannoehr et al., 2007; Wieler et al., 2011). Increased antimicrobial resistance development and spread in companion animals due to irrational antimicrobial usage, especially overprescribed broad spectrum antimicrobials without precise diagnostics, inevitably cause (1) animal health problem (increased mortality and morbidity), (2) economical problem to the owner (more visits-therapies and prolonged hospitalization), (3) economical problem to the veterinarian (possible loss of customers and high costs for hospital decontamination), and (4) human health problems (risks of zoonotic transmission). Because of this threat small animal veterinarians should prescribe broad spectrum antimicrobials after culturing and educate pet owners to handle infected-antimicrobial treated animals with precaution.

However, emergences of resistance toward antimicrobials which are critically important for human therapy are the most worrisome. These include the recent emergence of ESBL producing and carbapenemase positive Enterobacteriaceae bacteria in animal production (Horton et al., 2011), the emergence of farm associated MRSA ST398 (the main pig associated clone) (Cuny et al., 2010; Kluytmans, 2010; Weese, 2010) and of plasmid-mediated quinolone resistance in animal isolates and food products (Poirel et al., 2005; Nordmann et al., 2011). Unfortunately, there are several examples in the literature that show that these are already widespread in Europe and other parts of the world and have a large impact on human health (Angulo et al., 2004; Heuer et al., 2009; Forsberg et al., 2012).

Aquaculture (fish, shellfish, and shrimp farming industries) has developed rapidly in the last decade and has become an important food source (FAO, 2010). Fish pathogenic bacteria often produce devastating infections in fish farms where dense populations of fish are intensively reared. Although modern fish farming relies increasingly on vaccination and improved management to avoid infections (Markestad and Grave, 1997; Midtlyng et al., 2011), still many bacterial infections in fish are regularly treated with antimicrobials in medicated feed or by bath immersion. The most widely used drugs are fluoroquinolones, florfenicol, oxytetracyclines amoxicillin and sulfonamides (Cabello, 2006; Gräslund et al., 2003; Holmström et al., 2003; Primaver, 2006; Soothonchaikul and Garelick, 2009). By now, most of the fish pathogenic bacteria from fish farms with a history of infections have developed AMR (Colquhoun et al., 2007; Lie, 2008; Sørum, 2008; Farmed Fish Health Report, 2010; Shah et al., 2012a). Furthermore, in some areas of the world, particularly in South-East Asia, integrated farming is a common practice where organic wastes from poultry and livestock are widely used in manuring the fish farms (Hoa et al., 2011; Shah et al., 2012b). It has been reported that antimicrobial residues present in the poultry and livestock waste has provided sufficient selection pressure for the selection of AMR genes,
increasing the complexity of transmission of bacteria, and resist-
tances between the livestock and aquatic environment (Petersen
et al., 2002; Agersø and Petersen, 2007; Hoa et al., 2011; Shah
et al., 2012b).

AMR has been detected in different aquatic environments and
some resistance determinants have been found to originate
from aquatic bacteria. A good example is the recently emerging
plasmid-mediated quinolone resistance determinants from the
qnr family (Ash et al., 2002; Picao et al., 2008) and CTX-M from
aquatic Kluvyera spp. (Decousser et al., 2001; Rodriguez et al.,
2004; Ma et al., 2012). In addition, epidemiological and molecular
data indicate that some fish pathogens such as Aeromonas are able
to transmit and share AMR determinants with bacteria isolated
from humans such as E. coli (Rhodes et al., 2000; Serum, 2006).
Similarly, the fish pathogen Yersinia ruckeri have been reported to
share AMR plasmid and AMR genes with the bacterium causing
human plague (Welch et al., 2007).

ANITMICROBIAL USE IN PLANT AGRICULTURE
Streptomycin and oxytetracycline are routinely used for the pro-
phylaxis of fire blight disease (causative agent Erwinia amylovora)
in apple and pear orchards. Streptomycin use is strictly controlled
within the EU and is only authorized for use on a yearly basis.
However, streptomycin use in plant agriculture in the USA has
been replaced by oxytetracycline, due to streptomycin resistance
development among Erwinia amylovora in the apple orchards. Oxolinic
acid had been reported to be used in Israel against fire blight and
against rice blight in Japan (Shitenberg et al., 2001). Gentamicin
is used in Mexico and Central America to control Fire Blight and
various diseases of vegetable crops (Stockwell and Duffy, 2012).
However, the role of antimicrobial use on plants is, knowing the
AMR crisis in human medicine, the subject of debate (McManus
et al., 2002).

DISSEMINATION OF ANTIMICROBIAL-RESISTANT BACTERIA
THROUGH FOOD AND FOOD PRODUCTION
Resistant bacteria can be transferred from animals and plants
to humans in many different ways, which can be catego-
rized into three major modes of transmission: (1) through the
food chain to people (Roe and Pillai, 2003; Sootnorphitchai
and Garelick, 2009), (2) through direct or indirect contact with
livestock industry or animal health workers (Levy et al.,
1976), (3) through environments which are contaminated
with manure in agriculture (http://ec.europa.eu/environment/
integration/research/newsalert/pdf/279ana4.pdf) and aquaculture
(Petersen et al., 2002; Shah et al., 2012b). The environment con-
tains a great variety of bacteria creating an immense pool of AMR
genes that are available for transfer to bacteria that cause human
disease (Riesenfeld et al., 2004b; D’Costa et al., 2006). The real-
ization of these links sparked the recent interest in the role and
dynamics of environmental AMR (Figure 2).

In addition, other sources are available. For instance, wild
animals may also be carriers of antimicrobial-resistant bacte-
rria (Literak et al., 2011). These animals may have close contact
to human or farming areas and/or waste and become colo-
nized with resistant strains (Literak et al., 2011; Nkogwe et al.,
2011). Interestingly, animals in remote areas have been found to
harbor-resistant bacteria (Zhang et al., 2009; Glad et al., 2010;
Lang et al., 2010).

ANTIMICROBIAL RESISTANCE IN THE ENVIRONMENT
MICROBIAL COMMUNITIES IN SOIL AND ANTIMICROBIAL
RESISTANCE
Research data shows that in diverse soils from various regions of
the world, there is a wide dispersion of AMR. One explanation
for this phenomenon is the existence of antimicrobial produc-
bacteria in soil. The Actinomycetes, which are common
soil bacteria (Streptomycetes, Micromonospora, Saccharopolyspora
genus), synthesize over half of all known most clinically rele-
ant antimicrobials e.g., tetracycline, gentamicin, erythromycin,
streptomycin, vancomycin, and amphotericin. Bacteria of the
genus Bacillus also produce antibiotics, e.g., Bacillus brevis which
producing gamicidin (Balz, 2007). These antimicrobials now
also reach the environment from human and animal therapeu-
tics, through manure, sewage, agriculture, etc. Many retrospective
and prospective studies have been performed to assess the emer-
gence and selection of AMR in environmental bacteria. The
environment is eventually the largest and most ancient reservoir
of potential AMR, constituting the environmental “resistome”
(Aminov and Mackie, 2007; Allen et al., 2010; D’Costa et al.,
2011). Under such powerful selection pressure, it is not surpris-
ing that the soil resistome is so diverse (Knapp et al., 2010).
The best example illustrating this is the tetracycline resistome.
Tetracyclines are an important class of antimicrobials with desir-
able broad-spectrum activity against numerous pathogens and
the widespread emergence of resistance has had a massive impact
on these drugs (Thaker et al., 2010). Opportunistic pathogens
ubiquitous in the soil for example, Pseudomonas aeruginosa,
Acinetobacter spp., Burkholderia spp., and Stenotrophomonas spp.
can combine intrinsic resistance to several antimicrobials with a
remarkable capacity to acquire new resistance genes (Popowska
et al., 2010). Still, little is known about the diversity, distribution,
and origins of resistance genes, particularly among the as yet non-
cultivable environmental bacteria. In uncultured soil bacteria,
identified resistance mechanisms comprise efflux of tetracycline
and inactivation of aminoglycoside antimicrobials by phosphory-
lation and acetylation (Popowska et al., 2012). In addition, bacte-ia resistant to macrolides including the new drug telithromycin
have been reported from soil (Riesenfeld et al., 2004a). In a study
by (D’Costa et al., 2006), 480 strains of Streptomyces from soil
were screened against 21 antimicrobials. Most strains were found
to be multi-drug resistant to seven or eight antimicrobials on
average, with two strains being resistant to 15 of the 21 drugs. It
was also reported that soil is a reservoir for β-lactamases and these
genes, if transferred to pathogens, can then impact human health
(Allen et al., 2010). It is supposed that the presence of antibiotics
in the environment has promoted the acquisition or independent
evolution of highly specific resistance elements. These determi-
nants are located mainly on mobile genetic elements such as plasmids
and conjugative transposons, which ensure their spread
by horizontal gene transfer. Conjugative, broad-host-range plas-
mids play a key role in this process (Martinez, 2009; Stokes
and Gillings, 2011). Numerous studies have demonstrated that
the prevalence of such resistance plasmids in soil is very high
(Götz et al., 1996). Among the plasmids conferring resistance to antimicrobials, representatives of the incompatibility groups P, Q, N, and W have been identified. An example of this type of mobile genetic elements may be the IncP-1 plasmids (Popowska and Krawczyk-Balska, 2013). Results from the scientific literature show that plasmids carrying resistance genes have been identified in pathogenic bacteria of the genus Escherichia, Salmonella, Shigella, Klebsiella, Aeromonas, and Pseudomonas, the genera that can be found in soil and water (Stokes and Gillings, 2011). These plasmids carry determinants for resistance to at least one heavy metal (Ni, Cd, Co, Cu, Hg, Pb, Zn) and antimicrobials of different groups, i.e., tetracyclines, quinolones, aminoglycosides, sulfonamides, β-lactams, and chemotherapeutics (Sen et al., 2011; Seiler and Berendonk, 2012). Overall these data indicate that soil bacteria constitute a reservoir of resistance determinants that can be mobilized into the microbial community including pathogenic bacteria. Recent studies also indicate a different mechanism of AMR in soil-derived actinomycetes, by engendering mutations in genes encoding the transcriptional and translational apparatus that lead to alterations in global metabolism. This vertically selected AMR includes increased production of secondary metabolites (Derevecz et al., 2013). Very recently evidence for recent exchange of AMR genes between environmental bacteria and clinical pathogens was presented using a high-throughput functional metagenomic approach (Forsberg et al., 2012). In this study it was shown that multidrug-resistant soil bacteria contain resistance gene cassettes against five classes of antimicrobials (β-lactams, aminoglycosides, amphenicols, sulfonamides, and tetracyclines) with high nucleotide identity to genes from diverse human pathogens. Therefore, it is important to study this reservoir, which may contribute to the detection of new clinically relevant AMR-mechanisms and/or the multidrug-resistant pathogens that should be avoided from entering medically important bacteria (Torres-Cortés et al., 2011).

ANTIMICROBIAL RESISTANCE IN AQUATIC ENVIRONMENTS

Water is one of the most important habitats for bacteria, holding complex microbial communities. Not surprisingly, water also contains AMR bacteria. From natural fresh water systems to drinking water, or from sewage to human-engineered water infrastructures, AMR, either intrinsic or acquired, have been reported in aquatic environments worldwide (e.g., Goñi-Urriza et al., 2000; Volkmann et al., 2004; Schwartz et al., 2006; Ferreira da Silva et al., 2007; Böckelmann et al., 2009; Vaz-Moreira et al., 2011; Falcone-Dias et al., 2012). In this respect, given their characteristics, wastewater habitats are particularly important.

WASTEWATER HABITATS AS RESERVOIRS OF ANTIMICROBIAL RESISTANCE

Among the aquatic environments, wastewater habitats represent the most important reservoir of AMR bacteria and genes. This type of water contains human and animal excretions with abundant doses of commensal and pathogenic antimicrobial-resistant bacteria (Yang et al., 2011; Ye and Zhang, 2011, 2013; Novo et al., 2013). Since antimicrobials are not fully degraded in the human and animal body, antimicrobial compounds, their metabolites and transformation products are abundant in urban sewage treatment plants (Segura et al., 2009; Michael et al., 2013). Although proportion of the antimicrobial compounds are transformed and degraded in the environment, the occurrence of these micropollutants is reported worldwide, with antimicrobials of all classes being detected in wastewater habitats in concentrations ranging from ng⁻¹ to mgL⁻¹ (Michael et al., 2013). Simultaneously, urban sewage and wastewater contain AMR bacteria and other pollutants, such as pharmaceutical and personal hygiene products and heavy metals, whose effects on AMR selection are still not very clear (Graham et al., 2011; Oberlé et al., 2012; Patra et al., 2012; Novo et al., 2013). Often, wastewater treatment is not enough to eliminate the antimicrobial residues entering the system (Michael et al., 2013). The consequence is that such micropollutants, exerting selective pressures, may facilitate the selection of AMR bacteria or the acquisition of resistance genes by horizontal gene transfer (Martinez, 2009). Indeed, the relevance of wastewater habitats to the dissemination of AMR among human pathogens as well as commensal and environmental bacteria is increasingly emphasized (Baquero et al., 2008; Marshall and Levy, 2011; Czekalski et al., 2012; Rizzo et al., 2013). Wastewater treatment plants reduce the load of AMR bacteria, but treated water still carries elevated levels of AMR bacteria, and may select for strains with high levels of multidrug-resistance (Czekalski et al., 2012). Resistance gene abundance in a stream system could be linked to the input of (treated) wastewater and animal husbandry, demonstrating landscape-scale pollution of natural aquatic systems with AMR (Pruden et al., 2012). The currently available literature demonstrates that most of the AMR genetic elements found in clinical isolates are also detected in wastewater habitats, even shortly after they have been reported in hospitals (Szczepanowski et al., 2009; Rizzo et al., 2013). The occurrence of the same AMR genetic elements in different habitats demonstrates the uniqueness of the resistome, mainly due to rapid dissemination processes, demonstrating the urgent needs for an integrated approach.

Ubiquitous bacteria that can live in the environment and are also able to colonize humans are particularly relevant to the spread of AMR in the environment and the implications to human health. Indeed, numerous studies have reported the occurrence of AMR in ubiquitous bacteria isolated from wastewater habitats, which are also recognized as opportunistic pathogens, mainly nosocomial agents. AMR bacteria of clinical relevance which may be found in the environment comprise, among others, members of the genera Acinetobacter, Enterococcus, Escherichia, Klebsiella, Pseudomonas, and Shigella (Blanch et al., 2003; Reinthaler et al., 2003; Ferreira da Silva et al., 2006, 2007; Watkinson et al., 2007; Novo and Manaia, 2010; Czekalski et al., 2012). In addition, non-cultivable bacteria may also be important either for AMR spread or selection. Indeed, over the last years, the use of culture independent approaches brought additional insights into the abundance and diversity of resistance genes in wastewaters and into the effects of antimicrobials on the bacterial communities (Volkmann et al., 2004; Czekalski et al., 2012; Oberlé et al., 2012; Novo et al., 2013). In particular, several studies presented evidence that in wastewater habitats there is a high potential for horizontal gene transfer, mediated by plasmids and facilitated by integrons (Tennstedt et al., 2003; Szczepanowski...
et al., 2008; Moura et al., 2010; Zhang et al., 2011). Despite the importance of wastewater as a reservoir for AMR genes, and the relevance of wastewater treatment to control resistance spread, to date the number of studies that have been published remains relatively low (Figure 2).

Nevertheless, over the last decades the knowledge in this area has increased considerably and the importance of wastewater treatment systems for the spread of AMR is unequivocally demonstrated. Therefore, it is now possible to address some specific questions which we expect to be a focus of the research in this area in the coming years. Examples of these issues are (1) the identification of the conditions that may enhance or mitigate the occurrence of horizontal gene transfer and selection of AMR (which pollutants, which concentrations, temperature, pH, hydraulic residence time of wastewater treatment, etc.); (2) the classification and quantification of risk, e.g., the likelihood that an AMR bacterium or gene from wastewater habitats reach humans and causes issues for human health; (3) the improvement of wastewater treatment processes in order to minimize the loads of antimicrobial-resistant bacteria and genes in the final effluent (Dodd, 2012).

THE ANTIMICROBIAL RESISTANCE GENE POOL
AMR genes can be differentiated depending on the genetic event that is required for acquiring an AMR phenotype. These include genes that are acquired by horizontal gene transfer and genes that are present in the bacterial genome and that can encode AMR following gene mutations or activation (Olliver et al., 2005).

AMR features evolve as a consequence of permanent exchange of and ever new recombinations of genes, genetic platforms, and genetic vectors. Many of these genes are not primarily resistance genes, but belong to the hidden resistome, the set of genes able to be converted into AMR genes (Baquero et al., 2009). As evidenced by our discussion above, microbial organisms harboring these genes are present naturally in all kinds of environments, but also released into water and soil from organisms, including humans, where they evolve or increase in abundance under direct selection from exposure to antimicrobials. At the same time, antimicrobials (often at low concentrations), disinfectants, and heavy metals are disseminated into the water as well, and may act as selective factors fostering the evolution of new AMR features (Cantas et al., 2012a,b,c; Cantas et al., unpublished). The rate of degradation of antimicrobials in the environment varies and is dependent on a range of environmental conditions, for example: temperature, available oxygen, pH, presence of alternative sources of organic and inorganic discharges as described in Table 2.

HOW TO SLOW DOWN THE SPREAD AND EVOLUTION OF AMR?
In this review we have emphasized that the problem of AMR evolution and dissemination is multifaceted and involves clinical, agricultural, technical, and environmental systems. Similarly strategies to deal with the impending AMR crisis have to take this complexity into account.

The overuse of antimicrobials needs to be limited or reduced in human medicine, veterinary medicine, agriculture, and aquaculture. Ideally, the use of antimicrobials in agriculture should be eliminated. Intensive programs to educate both patients and physicians in reducing antimicrobial overuse should be implemented. Following the analysis more than 500 scientific articles, it has been suggested that the elimination of non-therapeutic use of antimicrobials in food animals, will lower the burden of AMR in the environment, with consequent benefits to human and animal health (FAAIR Scientific, 2002; Swartz, 2002).

Better management techniques and strict legislation in the use of antimicrobials for therapeutic use in humans and in animals will reduce the risk of development of AMR (Cunha, 2002; Defoirdt et al., 2011; Midtlyng et al., 2011). For example, the prevention of nosocomial transmission of multi-drug resistant bacteria is possible with active routine surveillance programs that can identify colonized patients. Numerous studies have demonstrated that such a “search and containment” approach and/or a “search and destroy” approach in which an attempt is made to eliminate carriage of the organism can reduce the incidence of hospital-acquired infections and be cost-saving (Muto et al., 2003).

New management techniques in the animal husbandry, such as organic farming, need to be thoroughly investigated to ensure that these are viable alternatives that help to reduce the potential for selection of AMR bacteria. Samples from organically farmed poultry showed a significantly lower level of AMR in intestinal bacteria such as E. coli and Campylobacter (Soonthornchaikul et al., 2006). However from organically farmed cattle no significant differences were obtained in microbiological contamination. E. coli and S. aureus isolates were found to have significantly lower rates of AMR in organically raised cattle (Sato et al., 2005). More studies are needed (1) to determine the reasons of antimicrobial usage in the farms by veterinarians, (2) to compare and update the recommended treatment protocols for veterinarians throughout different countries, (3) to evaluate the impact of other factors other than AMR development in bacteria: e.g., immune response-stress has been indicated to correlate with resistance genetic element shuffling among gut microbiota in different animal models.

### Table 2 | Degradation rates of various antimicrobials in soil.

<table>
<thead>
<tr>
<th>Class of antimicrobial</th>
<th>Degradation [%]</th>
<th>Time [d]</th>
<th>References</th>
</tr>
</thead>
<tbody>
<tr>
<td>Macrolides</td>
<td>0–50</td>
<td>5–30</td>
<td>Thiele-Bruhn, 2003*</td>
</tr>
<tr>
<td>Sulfonamides</td>
<td>0–50</td>
<td>22–64</td>
<td>Thiele-Bruhn, 2003</td>
</tr>
<tr>
<td>Fluoroquinolones</td>
<td>0–30</td>
<td>56–80</td>
<td>Hektoen et al., 1995; Thiele-Bruhn, 2003</td>
</tr>
<tr>
<td>Tetracycline</td>
<td>0–50</td>
<td>10–180</td>
<td>Björklund et al., 1991; Thiele-Bruhn, 2003</td>
</tr>
<tr>
<td>Aminoglycosides</td>
<td>0</td>
<td>30</td>
<td>Thiele-Bruhn, 2003</td>
</tr>
<tr>
<td>β-lactams</td>
<td>0–50</td>
<td>30</td>
<td>Thiele-Bruhn, 2003</td>
</tr>
<tr>
<td>Imidazoles</td>
<td>50</td>
<td>14–75</td>
<td>Thiele-Bruhn, 2003</td>
</tr>
<tr>
<td>Polypeptides</td>
<td>12–90</td>
<td>2–173</td>
<td>Thiele-Bruhn, 2003</td>
</tr>
</tbody>
</table>

* This reference does not include the modern macrolides with very long elimination half-lives. For instance, Tularomycin has an half live (so 50% degraded, not nearly 100%) in soil of 99 days (Pfizer, personal communication 2013).
such as: Atlantic salmon, zebrafish, neonatal piglets, and cats (Cantas et al., 2011, 2012a,b,c, 2013). Animal welfare parameters under intensive production such as stress should be investigated in future studies with regards to control of resistance development in animal husbandry.

Vaccination and improved hygienic measures are among the important cornerstone in controlling infectious diseases and consequently aid in reducing AMR (Potter et al., 2008). The Norwegian aquaculture may serve as a good example by reduction in the use of antimicrobials from around 50 tons in the late 1980’s to less than 1000 kg per annum after introduction of effective vaccines against devastating fish diseases like furunculosis and vibriosis (Midtlyng et al., 2011; NORM/NORM-VET, 2011).

The use of pre- and probiotics to improve the health and performance of livestock might be a good alternative to growth promoters. This is an important biological control aiming to reduce outbreaks of infectious diseases and which in turn would minimize the use of antimicrobials in livestock and aquaculture for therapeutic purposes (Verschuere et al., 2000; Callaway et al., 2008).

The issue of dissemination and possible long-term enrichment of AMR and AMR genes in the environment (Knapp et al., 2008) needs to be studied further, with specific regards to the actual risks associated with it. However, taking action is already possible today. For example, several treatment methods for waste and wastewater disinfection and removal of micropollutants, including antimicrobials, are available. These include various chemical disinfections, UV treatment, and membrane filtration. Disinfection and DNA degradation of community based and hospital wastewater may be effective means to reduce AMR release, although more research is required to fully assess the inactivation of resistance genes (i.e., DNA released from lysed cells that may be available for horizontal gene transfer) by these measures (Dodd, 2012). The combined removal of pollutants that are potential selective agents, disinfection, and deactivation of the genetic material, may be a useful strategy to reduce the pollution of environments with resistance factors.

CONCLUSIONS

Inevitably, AMR in medicine has become common place. Bacteria have evolved multiple mechanisms for the efficient evolution and spread of AMR. Meanwhile the new developments of quick and adequate molecular diagnostic techniques for the identification and epidemiological surveillance of genetic determinants of AMR in different hosts and in the environment will enhance the number of control options. We have outlined above a number of potential measures that are enabled by our improved ability to track AMR. However, there seems to be a clear need for action and policy changes. This includes drug licensing, financial incentives, penalties, and ban or restriction on use of certain drugs. Similarly, the prescriber behavior needs to be altered. Animal health and hygiene needs to be improved. In addition, the implementation of microbiological criteria for the detection of certain types of resistant pathogens would be important to control the trade of both food animals and food products. The problem of AMR in human medicine will not be solved if nothing is done to limit the constant influx of resistance genes into the human microbiota via the food chain or contact with the environment. Introduction of antimicrobial compounds into the aquatic environment via medical therapy, agriculture, animal husbandry and companion animals has resulted in selective pressures on resident environmental bacteria. Development of AMR in environmental bacteria has a great impact and may help in explaining how human and animal pathogens acquire resistance features. Besides the role of clinical microbiology laboratories with rapid and accurate detection of a diverse number of pathogens and its drug resistance profiles, robust routine surveillances in an epidemiological frame-work covering the whole livestock “food chain” and the environment need to be taken into consideration. Due to this complexity the control of AMR has to include numerous actions at diverse levels. Future research should focus on finding unknown routes of transfer of AMR between microbiotas of relevance to the food chain and to all microbiotas of importance for bacterial pathogens when they acquire antibiotic resistance genes laterally. Ultimately, even a successful integrative approach on all aspects, can probably only help to slow down the spread of AMR, not prevent it. The development of new generations of antimicrobial should therefore receive equal attention. This is summarized and emphasized in a 12-point action plan against the rising threats of AMR implemented by the European Commission which includes actions in the field of human medicine, animal husbandry, veterinary medicine, authorization requirements for commercialization of human and veterinary drugs and other products, on research, on scientific opinions, and undertaking also actions on the international level in collaboration with the WHO and Codex (http://ec.europa.eu/dgs/health_consumer/does/communication_amr_2011_748_en.pdf, Bush et al., 2011).

The problem of AMR is widespread all over the world, therefore it is not eradicable, but can be managed. Concerted efforts between medical doctors, dentists, veterinarians, scientists, funders, industry, regulators, and multi-disciplinary approaches are needed to track resistance. Furthermore, global monitoring of the antimicrobial drug consumption in human and veterinary medicine and AMR, is an essential part of an overall strategy to inform, educate and get commitment of all parties, including farmers and patients (American Academy of Microbiology, 2009). All these are important measures for the efficient future use of antimicrobials in medicine. All members of society should be conscious of their role and take on responsibility for maintaining the effectiveness of current and future antimicrobials. We believe that future interventions can be successful in minimizing this problem.

AUTHOR CONTRIBUTIONS

L. Cantas defined the review theme, established the interdisciplinary coordination and the collaborations, designed the manuscript, contributed to the data collection, data analysis, and drafting and writing of the manuscript. Syed Q. A. Shah, L. M. Cavaco, C. M. Manaia, F. Walsh, M. Popowska: drafting, writing, and editing the manuscript. H. Garelick and H. Bürgmann: contributed to data analysis, drafting, and writing the manuscript. H. Serum: contributed to manuscript design, data collection, data analysis, drafting, and writing the manuscript. All authors have contributed to, seen and approved the manuscript.
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