Food is a basic requirement for life, providing the essential nutrients and energy required for optimal health. However, food may also be associated with adverse health effects, because it may contain natural toxins, hazardous chemical substances or pathogenic microorganisms that can affect health negatively. Additionally, it is possible that the dietary intake of specific nutrients in foods is either too low or too high, resulting in potential deficiencies or toxicity symptoms.

The diverse causes of these health effects associated with food consumption and the demand for advice on safe and healthy diets have led to the development of different research disciplines in food safety and nutrition. The negative health impact of human exposure to chemical substances and pathogenic microorganisms through food is evaluated in two separate disciplines, chemical and microbiological risk assessment. Apart from that, both health risks and health benefits associated with foods and diets have been studied through the discipline of nutrition. However, in the past decade, the joint assessment of risks and benefits associated to hazardous agents, food compounds, nutrients, single foods and whole diets has been taken up, resulting in the establishment of “risk-benefit assessment” (RBA) as a new multi-disciplinary and integrated scientific discipline (Boué, Guillou, Antignac, Bizec, & Membré, 2015; Tijhuis et al., 2012; Verhagen et al., 2012a).

With the overall aim of exploring how RBA can be further developed, this paper aims to describe the recent progress in RBAs and to identify and discuss key challenges in RBA research. To clarify the fundamentals of RBA and to provide a basic understanding of the background of many of the challenges, the main concepts of the underlying disciplines chemical risk assessment, microbiological risk assessment and nutritional risk and benefit assessment are explained. Following that, the developments in RBA thus far are addressed. The major part of the paper is devoted to a discussion of ten challenges, as well as to suggestions for how they can be met. The conclusion
summarizes the authors’ vision on the future developments of the research area.

1.1. Risk and benefit assessment in food safety and nutrition

The use of risk assessments has traditionally been an integrated part of a common risk analysis framework (Fig. 1), where risk assessment is done by risk assessors who provide scientific advice to support decision making by risk managers, such as food authorities or food producers, on the potential risks associated with food consumption. Risk communication is an essential part of the risk analysis, both between risk assessors and risk managers, and between assessors, managers and other stakeholders (FAO/WHO, 2006a).

Risk assessment was first formalised for chemicals by the establishment in 1980 of the International Programme on Chemical Safety (IPCS), which proposed a scientifically based process including four elements: hazard identification, hazard characterization, exposure assessment and risk characterization (Fig. 2). The first step, hazard identification, involves the identification of the inherent toxicological properties of a chemical substance in the food that may affect human health adversely. Depending on the nature of the chemical substance, the information on hazards may stem from in vitro studies (for example on genotoxicity), experimental animal studies, and human data. The next step, hazard characterization, involves dose-response evaluations of the toxicological effects of the chemical substance that are identified in the previous step, including identification of critical effect levels such as no observed adverse effect level (NOAEL), lowest observed adverse effect level (LOAEL), or a benchmark dose (BMD) (IPCS, 2010). These critical effect levels are based on either acute or chronic effects and are usually determined on the basis of results obtained from animal experiments. After applying uncertainty factors to account for differences in sensitivity between species (e.g., animal to man) and within the human population, the critical effect levels are translated to health-based guidance values such as acceptable daily intake (ADI), tolerable daily intake (TDI) or acute reference dose (ARfD) (IPCS, 2010). In exposure assessment of the chemical substance, the exposure from food is estimated by use of accurate and representative data of relevant food consumption and occurrence of chemical substances in the foods. The last step, risk characterization, integrates the outcomes of the hazard characterization and the exposure assessment, and the output is given to the risk managers.

Microbiological risk assessment has mainly been used for bacterial pathogens, but it has also been applied to viruses and parasites. It was developed after chemical risk assessment was established and adopted much of the terminology. However, the nature of microorganisms has led to specific challenges, which resulted in some essential differences in the definitions (see Section 2.1), as well as in the risk assessment methodology (Lammerding, 2013).

First, the definition and identification of the microbiological hazard are complicated by the fact that microorganisms adapt and evolve over time, so new strains can emerge with different characteristics than those that were originally described. Next, the dose-response relation typically describes acute health effects, with the probability of acute illness being described as a function of the ingested dose in a single meal. Due to the differences in responses between humans and animals, data for microbiological dose-response models can usually not be derived from animal experiments. As an alternative, human data are required, but these are not easily obtained. The use of biologically plausible “single hit” models that assume that, with low probability, a single bacterial cell can lead to illness, is a general practice in microbiological dose-response modelling (Haas, Rose, & Gerba, 2014; FAO/WHO, 2003). Exposure assessment is complicated by the fact that living organisms can multiply, and consequently, the occurrence of microbial growth and inactivation imply that concentrations can change during food processing and storage. Therefore, concentration data alone are insufficient and the ingested doses have to be estimated by means of mathematical modelling in so called "process risk models" that apply predictive models for growth and inactivation (FAO/WHO, 2008; Zwietering & Nauta, 2007). Note that this implies that, in contrast to chemicals, exposure depends on the growth and inactivation characteristics of the microorganism of concern (Fig. 2). Critical limits for the presence of microorganisms are generally not determined on the basis of the hazard characterization only, so equivalents of NOAEL and BMD as used in toxicology are not applied. Instead, risk-based microbiological targets such as food safety objective (FSO) are used, which are derived from risk characterization, i.e., a combination of hazard...

Risk assessment of essential nutrients follows the same principles as chemical risk assessment, with the notion that essential nutrients have a dual risk relationship with risks occurring at both the upper end (excess) and lower end (deficiency) of the intake range (NCM, 2014). Another distinct feature is that data on adverse effects in relation to excessive or deficient amounts of nutrients are often available from human studies, which compared with chemical risk assessments overall, may reduce the size of uncertainty factors applied. The tolerable upper intake level (UL) is therefore the maximum level of chronic daily nutrient intake from all sources judged to be unlikely to pose a risk of adverse health effects to humans (EFSA, 2006) and thus includes an uncertainty factor as in the case of chemicals. The lower threshold intake (LTI) is the level of intake below which, on the basis of current knowledge, almost all individuals will be unable to maintain “metabolic integrity”, according to the criterion chosen for each nutrient (EFSA, 2010b).

Consideration of specific nutrient intakes associated with adverse health effects above or below specific intake levels has received less attention in the nutrition area compared with non-nutrients, such as drugs, food additives, and pesticides (IOM, 2007). The concept of the risk assessment of nutrients was stimulated by the IPCS in 2002 (IPCS, 2004), and by the Codex Alimentarius, FAO/WHO, EFSA and others (FAO/WHO, 2006c; Aggett, 2010; Taylor & Yetleya, 2008). In addition, the implementation of an organized nutritional risk assessment approach for scientific reviews has been stimulated by the increased use of food supplements, fortified- and functional foods and subsequent requests by regulatory agencies to identify upper levels of nutrient intake (Taylor & Yetleya, 2008; Taylor, 2007). In 2010, EFSA published a scientific opinion on the general principles for development and application of dietary reference values (DRV) (EFSA, 2010b), and other DRV processes have followed the same risk assessment approach, including the update of the Nordic Nutrition Recommendations (NCM, 2014).

Current approaches thus predict a threshold above which the nutrient intake is excessive, and another threshold below which the intake is inadequate, while an intake range between these two boundaries can be considered an ‘optimal’ intake range within which the recommended intake and the benefit assessment is set (NCM, 2014). Nutritional benefit assessment may thus be considered as the intake range beyond which there is a risk. Nutritional RBA can be broadened to not only consider nutrients, but also to include any excess or deficient intake of foods, diets or energy.

One example of the application of benefit analyses is the European health claim regulation, which states that health claims should be “substantiated by generally accepted scientific evidence and by taking into account the totality of the available scientific data, and by weighing the evidence” (EU Commission, 2006). The steps involved in the assessment of health claims include identification and characterization of the food or the food compound, definition of the effect and assessment of whether such an effect can be considered beneficial to human health. Finally, the scientific substantiation for a beneficial effect is assessed based on the totality of the current evidence between the consumption of the food or the food compound and the claimed effect studied in the appropriate target group (EU Commission, 2006).

A comparison of the application of risk and benefit assessment for chemical substances, microorganisms and nutrients shows that, traditionally, risks are considered for all, but benefits only in nutrition. An essential difference between different types of risk and benefit assessment is illustrated in Fig. 3. Typically, looking at both acute and chronic adverse effects, chemical and microbiological risk assessments investigate situations where exposure is to be considered “too high”. This implies that the risk increases with higher doses, and threshold doses may be derived as cut-off points below which the intake is considered safe, or the associated risk is considered acceptable (Barlow et al., 2015). In contrast, within nutrition, both the situation where there is a risk of nutritional deficiency and the situation where there is a risk of nutrient intoxication are relevant, creating a “window of benefit” (Palou, Pico, & Keijer, 2009; Tijhuis et al., 2012). Interestingly, research in situations where the intake is too high (above the upper intake level (UL)) is commonly referred to as toxicology, whereas research considering beneficial intake or too low intake, is part of nutrition.

### 1.2. The development of risk-benefit assessment

Although independent risk and benefit assessments have proven to be useful for decision support in food safety and nutrition, their results may be too much focused on one hazard, one food compound or one health effect. When establishing guidelines and advice on food consumption, nutrient intake and diet choices, there is a need for an overarching approach, in which all of the relevant health risks and benefits are included and compared. This need for RBAs has been identified earlier in several publications (EFSA, 2007; EFSA, 2010a; Renwick et al., 2004) and led to the development of RBA of foods as a new research discipline. An RBA is multidisciplinary by nature, and may require expertise from not only toxicologists, microbiologists, and nutritionists, but also from epidemiologists, chemists, librarians, statisticians, and medical scientists. As proposed in the EU-funded project BRAFO (Benefit-Risk Analysis of Food) (Boobis et al., 2013), it is common to use the risk analysis and risk assessment frameworks (Figs. 1 and 2) as the basis for the RBA methodology by applying the established concepts to both risks and benefits. A recent extensive review of studies related to the combined RBA of foods, nutrients and compounds shows that the majority of published studies have been related to fish consumption where the nutritional beneficial effects are compared with the adverse effects from chemicals (Boué et al., 2015). This RBA of fish (e.g., (Hoekstra, Hart et al., 2013)) is an example of an RBA case where the content of polyunsaturated fatty acids, and in particular docosahexaenoic (DHA), and eicosapentaenoic fatty acids (EPA), recognized for their health benefits, is counterbalanced by the content of pollutants such as methylmercury and dioxins, known to potentially induce adverse health effects. There is also an example of microbiological aspects being added to an RBA of fish (Berjia, Andersen, Hoekstra, Poulsen, & Nauta, 2012).

Several European projects have been conducted in which methods and modelling frameworks were developed, leading to considerable progress in the risk-benefit area (Boobis et al., 2013; Hart et al., 2013; Hoekstra et al., 2012; Verhagen et al., 2012a). Among others, the BRAFO project and EFSA developed the "tiered approach" to be used as a general framework for RBA1 (Fransen et al., 2010; Hoekstra et al., 2012). The basis is that a number of tiers have to be evaluated before making a decision on the required steps to be taken in the RBA. This approach proposes that a qualitative assessment is sufficient if data are scarce or there is clear evidence that risks outweigh the benefits (or vice versa). If the balance between benefits and risks is unclear, the assessment has to be performed at a higher tier, including quantitative assessment. As part of the BRAFO project, a number of relevant risk-benefit studies that illustrate the usefulness of a tiered approach for RBAs have been performed (Hoekstra et al., 2008; Schütte et al., 2012; Verhagen et al., 2012b; Watzl et al., 2012). A specific software tool, QALIBRA, has been developed to facilitate the performance of quantitative assessments in the final tier (Hart et al., 2013; Hoekstra, Fransen et al., 2013).

### 2. Challenges in risk-benefit assessment

Although significant progress has been made in the development of

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1 Within the BRAFO project, the term benefit-risk assessment was preferred over risk-benefit. For consistency we consequently use risk-benefit assessment (RBA) throughout this paper.
methods and terminology in RBA, several challenges remain. Some of these challenges relate to the differences between the underlying research disciplines, which have different use of terminology and different approaches for the assessment of health effects related to the consumption of food. Other challenges relate to the specific objective of RBAs, the scarcity of the required data, or the complexity of the characterization of health effects. Below, we provide a description of ten major challenges that were identified during the course of working with RBAs, with explanations of the challenges and discussion on the way forward for meeting them in the future.

2.1. Definitions

The different approaches used in the disciplines contributing to RBA (Section 1.1) apply different terminology or may apply the same terminology in a different way. Dissimilar definitions can lead to confusion and lack of understanding of the risk-benefit question (Section 2.3). As an example, the central concept of "hazard" is defined differently in various contexts. Published definitions of hazard include "inherent property of an agent or situation having the potential to cause adverse effects when an organism, system, or (sub)population is exposed to that agent" (IPCS, 2004), "the potential of a risk source to cause an adverse effect(s)/event(s)" (Renwick et al., 2003) and "a biological, chemical or physical agent in, or condition of, food with the potential to cause an adverse health effect" (CAC, 2011, p. 112). In the latter definition, the hazard is the agent (or risk source, that is the pathogen, chemical substance or food compound) and in the others it is an inherent property or the potential of this agent. Due to this difference in definitions, the hazard is usually synonymous to the pathogen(s) of concern in microbiological risk assessment, whereas it usually is the potential health effect caused by the chemical substance or food compound in chemical risk assessment and nutrition (Barlow et al., 2015).

Similarly, there are different definitions of "risk", for example "the probability of an adverse effect in an organism, system, or (sub)population in reaction to exposure to an agent" (EFSA, 2010a; IPCS, 2004), or "a function of the probability of an adverse health effect and the severity of that effect, consequential to a hazard(s) in food" (CAC, 2011). So in one definition the risk is a probability, in the other, it is a combination of probability and severity.

When mirroring risk assessment to benefit assessment, the benefit is defined at a level comparable to both the hazard and the risk (Boobis et al., 2013; EFSA, 2006), so "benefit" is both the counterpart of "hazard" and the counterpart of "risk". Hence, the term "benefit" can be used for anything between the agent causing the health effect and the probability and magnitude of that effect. Moreover, when used as equivalent of "risk", the benefit is not necessarily interpreted as the probability of a positive effect, but commonly as the decrease in the probability of an adverse health effect. This wide interpretation of the one of the central concepts in RBA can be considered confusing.

The present definitions can be well understood in a historical perspective, given that RBA has evolved from a variety of disciplines. However, for further development, the discipline "risk-benefit assessment of foods" needs a clearer set of definitions and harmonized terminology that is comprehensible for all those involved. To accommodate the fact that some agents or food compounds (i.e. "hazards" of "benefits") can both be a source of positive and negative health effect depending on the exposure (Fig. 3), Boué et al. (2015) propose to use the term "health effect contributing factor" (HECF) for "the agent able to cause an adverse or positive health effect in the case of exposure". This is a useful first step in the reconsideration of the terminology used in RBA. Consensus within the international research community is required for clarification and harmonization purposes and definitely when it would be used for regulatory purposes. Obtaining such a consensus is a process that should be led by international authorities, and should include representatives of all relevant disciplines involved in RBA.

2.2. Bottom-up versus top-down approach

In this paper, we distinguish between two overall approaches to assess health effects in RBA and refer to them as "bottom-up" and "top-down". This terminology is derived from studies in microbiological food safety aimed at ranking microbiological food risks (Cassinì et al., 2016; EFSA, 2015). The two approaches are characterised by their different starting point. The typical risk assessment approach, which starts with the hazard identification for the food product or its ingredients and finishes with the human health outcome obtained after combining the exposure assessment with a dose-response model (Fig. 2), is referred to as the bottom-up approach. The alternative top-down approach starts with the adverse (or beneficial) health outcomes as obtained from human observational studies, i.e., incidence data and identified risk factors. These are then traced back to the food sources that caused the disease of concern (or benefit of desire), thus linking the health effect to the food product.

A similar distinction in approaches can be made in nutritional and chemical risk assessment. The usual risk assessment approach (i.e., bottom -up) is targeted at intake of specific nutrients or food compounds, and the dose-response relation is typically derived from animal experimental data. The alternative top-down approach is an approach where relative risk estimates from human observational studies are used and linked to foods or food compounds that are identified as risk factors. In the review of the BRAFO project, Boobis et al. (2013) identify these two approaches as one based on experimental animal data (bottom-up) and one based on human observational studies (top-down). We prefer the bottom-up and top-down terminology as it is more generic and can also be applied for microbiological risk assessment, which does not apply animal data.

Hence, with the bottom-up approach, the assessment starts with the food product, food compound or contaminant, followed by an exposure assessment and a dose-response model used for the risk-benefit...
characterization. An advantage of this approach is a direct causal link between intake of the food product or food compound (or contaminant) of concern and the associated health effect. A disadvantage is that there may be a large uncertainty attending the exposure assessment and (especially) the dose-response.

With a top-down approach, the starting point of the analysis is the incidence of a health outcome in the consumer. Typically, data from epidemiological studies (case-control studies, cohort studies, randomized controlled trials) are used to associate human health outcomes with risk factors that are defined in terms of food consumption, allowing for the estimation of metrics such as the odds ratio or the relative risk. These measures of association are then combined with population statistics and incidence data to estimate the actual health risks in the population. The relative risks may also be used to construct a dose-response relation, where the relative risk is a function of the intake as specified in the underlying study. The strength of human observational studies is that they are based on actual health effects, measured in specified populations. Weaknesses are that the observed associations are not a proof of causation, that the studied population may not be representative for the population group of interest and that many data are required if the health effect of interest is small. For microbial pathogens, a top-down approach can be used to estimate the number of cases of disease caused by a pathogen due to its presence in a specific food, a method referred to as “source attribution” (Pires et al., 2009). Here incidence data on a specific health outcome (e.g., gastroenteritis caused by salmonellosis) is traced back to a specific food source (e.g., chicken meat) by the use of subtyping information of isolates of the pathogen in human cases and food sources.

Generally, within RBA, it is necessary to use different approaches for different health effects of food compounds or contaminants. For example, in the studies on fish of (Berjia et al., 2012; Hoekstra, Hart et al., 2013) (Fig. 4), the effects on coronary heart disease, stroke and neurological development of children (IQ) are derived from top-down approaches, but those related to exposure to dioxins and Listeria monocytogenes are derived from bottom-up approaches. The reason for the application of these different approaches is obviously the availability of data, which in turn is related to the feasibility of acquiring the requested data and also the quality of the studies providing the data. Still, if different approaches are used to obtain different health effect estimates in the same RBA, it may be hard to compare them. Not only can there be a difference in the known bias associated with the approach (such as a potential to overestimate the risk obtained from dose-response models derived from animal experiments), but also the nature of the uncertainties associated with the assumptions of the approaches will be different (Section 2.5).

Studies that combined and compared bottom-up and top-down approaches may help clarify how the two methods can be integrated in RBA. For example, Bouwknecht, Knol, van der Sluijs, and Evers (2014) compared the approaches in a case study on Campylobacter in the Netherlands and identified the differences in the underlying uncertainties. They found that the difference in the point estimates of the risks as found by the different approaches can be large, but they still have overlapping uncertainty intervals. This implies that one cannot a priori conclude that one approach is better than the other. It is advisable to aim for evidence synthesis by using an approach that takes advantage of all available data and combines bottom-up and top-down approaches. One option for evaluating such a combined approach is the performance of simulation studies where the expected results of a hypothetical epidemiological study are investigated on the basis of a risk assessment.

2.3. The risk-benefit question

The crucial initial step of an RBA is the definition of the risk-benefit question (Hoekstra et al., 2008) or problem definition (Boobis et al., 2013; Boué et al., 2015; EFSA, 2010a). The risk-benefit question is generally a comparison between two, or a series of, choices, alternative policies or courses of action, described in the form of scenarios (Boobis et al., 2013). In these scenarios, both positive and negative health effects have to be taken into consideration. When a series of scenarios is compared, the risk-benefit question can be used to identify the optimum intake (Berjia et al., 2014). An aim of the risk-benefit question is to specify the RBA-task in such a way that it is feasible and will provide useful results. For example, an RBA of fish should indicate what sort of fish (e.g., lean/fatty, farmed/wild), target population group, and in general any other constraint that could narrow the risk-benefit question. In the end, the level of specification of the question will also depend on the data available.

As a variety of risk-benefit questions can be asked, it can be helpful to categorise them and to identify specific approaches that can be used to answer these different categories of questions. Here, one type of categorisation is the level of aggregation: the risk-benefit question can be targeted at a food compound level (a nutrient, a chemical or microbiological contaminant), a food product level (e.g., fish) or a diet level (Hoekstra et al., 2008).

When the risk-benefit question is targeted at the food compound level, it should be a compound that is associated with both positive and negative health effects, e.g., a (macro-) nutrient. Examples for RBAs directed at the food compound are those for folic acid (Hoekstra et al., 2008) and vitamin D (Berjia et al., 2014) (Fig. 4). The choice between a bottom-up or top-down approach will depend on whether the health effects associated with food compounds are obtained from animal experiments or human observational studies. To assess the total intake of the food compound, it will be necessary to consider the intake of all relevant foods and food products in the diet that contain it, and the concentrations of the compound in these foods and food products have to be known. As this can be rather complicated, one can choose a risk-benefit question that only considers a difference in intake or concentration in one or a few food products, making some assumptions for the background diet.

When the risk-benefit question considers a food product, the positive and negative health effects can be associated with different food compounds or contaminants that it contains. Typical examples of RBAs directed at this level of aggregation are those performed for fish (Berjia et al., 2012; Hoekstra, Hart et al., 2013, Fig. 4). The health effects of the intake of the food product may be directly available from epidemiological data or a human trial study, allowing the use of a top-down approach. Relative risk estimates can inform about the health impact of one intake scenario compared with another. Alternatively, a bottom-up approach may be used where all relevant food compounds (and contaminants) in the food product have to be identified and comprised in the RBA to assure that the health effects of interest are included. In that case, a selection of relevant food compounds and contaminants needs to be made based on the associated levels of evidence and the precise risk-benefit question. However, because in some cases only exposure through the selected food product is considered, and not the total exposure from all food products containing the compounds, it is difficult to use a bottom-up approach with a dose-response relation for each compound.

When considering a whole diet, the bottom-up RBA approach will usually not be feasible, unless the risk-benefit question is clearly delimited: the number of food compounds (and contaminants) and their combined intakes easily get too large for a complete exposure assessment and hazard characterization. However, a top-down approach using studies on human consumption may be possible if the appropriate data are available, for example from a dietary intervention study. Van Kreyl, Knap, & Van Raaij, 2006, performed a study to analyse the health effects of the current diet in the Netherlands that may be regarded as an RBA of diets, but otherwise, to our knowledge, no formal RBAs of whole diets have been performed so far.

In each of these three categories of risk-benefit questions, the options for inclusion and exclusion of food compounds and contaminants,
food products and health effects are large. To clarify the selected elements in the risk-benefit question, we propose the use of schematic framing of the risk-benefit question, as exemplified in Fig. 4 for four published risk-benefit studies for food compounds or food products. A scheme like this is broadly applicable and may offer a transparent way to identify different types of risk-benefit questions and clarify how the risk-benefit question is addressed. In the case of an RBA of a whole diet, the scheme would be pretty complex, which stresses the difficulty of doing an RBA of a whole diet.

2.4. Lack of data and knowledge and the consequential uncertainties

The data needs for an RBA are large and diverse. RBAs frequently face data gaps and lack of knowledge, such as lack of human data, information on dose-response and intake levels for specific population groups. These challenges are also faced in other modelling exercises (such as many risk assessments), and need to be addressed by documentation and discussion of the assumptions made. A consequence of limited data and lack of knowledge is that the uncertainty related to the assessment may be large. Yet, characterising this uncertainty is crucial in the risk-benefit characterization.

As part of the QALIBRA project, Hart et al., 2013, provided an overview and discussion on the importance and challenges related to uncertainty in RBA and described strategies to deal with uncertainty. The QALIBRA software tool developed in the project allows the user to perform stochastic RBA and, as part of that, analyse uncertainty, either by quantitative methods or by qualitative scenario analyses. This has been an important step forward for the analysis of uncertainty within RBAs.

Still, as previously identified by Boobis et al., 2013, and others, there are different areas within RBAs where lacking data creates a major challenge. An important area is dose-response modelling. For chemical substances, the dose-response relations are usually derived from animal experimental data, where a set of assumptions is needed to establish a threshold that can be applied for human consumers. As the objective of these dose-response relations in animals is often to identify potentially dangerous doses and to set safe health-based guidance values such as the ADI or TDI, the assumptions may tend to overestimate the true human health risks. Yet, for RBAs, it is important to derive the magnitude of the positive and negative health effects in the same way and therefore one needs the best possible estimate of the likelihood of the health effect from a dose-response relationship, not the “worst case” value. For chemical dose-response relationships, this means that the use of BMD models may be preferred over NOAELs and LOAELs, and that the uncertainty factors used to translate animal data to human guidance values may not be appropriate if the dose-response relationship is to be applied in RBAs.

The uncertainty attending the dose-response relations for microbial pathogens is also large. These dose-response relations are usually based on human volunteer studies or outbreak data, which means they are based on limited data sets, for specific strains and specific population groups, and generalised thereafter. Dose-response relations based on studies with healthy young volunteers may be expected to underestimate the risk, whereas those derived from outbreaks (with more...
virulent strains) may overestimate the risk. Further, it is known that immunity plays an important role and may lead to overestimation of the risk, but it is difficult to include this in the modelling (Havelaar & Swart, 2014).

Another uncertain element of the dose-response modelling is the long-term effect of exposure, which is specifically relevant for chemical substances. An acute effect is the direct consequence of an individual ingested dose and therefore relatively easy to describe in dose-response model. For long-term effects, however, it is much harder to identify how different doses accumulate into health effects. The use of physiologically-based pharmacokinetic (PBPK) models (Boobis et al., 2013; Zeilmaker et al., 2013) can be useful, but these models still need further development.

If the dose-response modelling is based on relative risk estimates obtained from human observational studies, uncertainties may be large as well. Some important issues are, for example, the uncertainty regarding the causality of observed associations between risk factor and effect and the representativeness of the data. To account for the uncertainties, top-down approaches (using this type of effect modelling) and bottom-up approaches (using the other dose-response relations) may be combined in a comparative analysis (Section 2.2).

Uncertainties are an inevitable intrinsic element of science, risk assessment and RBAs, and it is of utmost importance that they are not ignored. A challenge here is that, as in risk assessment, it is not primarily the objective of an RBA to assure that the uncertainty is small enough (as aiming for a p-value smaller than 0.05), but to indicate how large the uncertainty actually is (Nauta, 2007). One should deal with the identified uncertainties by explicitly addressing and characterizing them in the assessment and by clearly communicating them to all stakeholders. By framing the risk-benefit question (Fig. 4) and addressing the required data, RBA models can be important in identifying the most important data gaps and the crucial lack of knowledge. Thus, they can guide future data generation and research. Setting the future research agenda based on the most important sources of uncertainty can therefore be one of the key outputs of an RBA.

2.5. The imbalance in level of scientific evidence

The level of scientific evidence needed for identifying negative and/or positive health effects of a food compound, food or diet is not consistent (Boobis et al., 2013), because the presence of benefits and the absence of risks need to be guaranteed (Hoeckstra, Hart et al., 2013; Tijhuis et al., 2012). In the case of health claims, a nutritional benefit needs to be scientifically substantiated with convincing evidence of the cause and effect relationship, before it can be accepted according to the current EU regulation (Section 1.1). At the other hand, in the case of setting dietary guidelines, a nutritional benefit of a food or food group may only need to be scientifically substantiated at the level of probable likelihood of an association (Kromhout, Spaatj, de Goede, & Weggemans, 2016; Tetens et al., 2013; WHO, 2003). Finally, the level of scientific evidence needed for identifying risks or negative health effects may be small, as only an indication of a risk is sufficient for the scientific substantiation.

Due to this discrepancy in the level of scientific evidence needed for considering a food compound or contaminant as a “hazard” or a “benefit”, risks are more likely to be included in an RBA than benefits, thus leading to a potential bias in the RBA (Boobis et al., 2013; Hoeckstra, Hart et al., 2013; Tijhuis et al., 2012). Another consequence of this discrepancy is that different types and levels of uncertainty will be associated to the risk assessment on the one hand and the benefit assessments on the other, which complicates the characterization of the combined RBA even further (Section 2.4).

The imbalance in the required level of scientific evidence for risks and benefits demands a paradigm shift from the RBA as a sum of risk and benefit assessment to the RBA as a well-integrated risk-benefit assessment. Such a well-integrated RBA deals not so much with studying a hypothesis about the presence or absence of a health effect associated with the intake of a certain amount of food product or food compound or contaminant, but predominantly with the size of the health effects. Even though the strength of evidence for the presence of a health effect is strongly correlated to the size of the effect, these are not the same thing. Stochastic modelling techniques, which include quantification of uncertainty and variability, allow an evaluation of potential health effects, even if the effects themselves are not statistically significant. In doing so, it may be possible to study how the estimated size of the effect, and some alternative scenarios about these effects, may impact public health. From this, one might conclude that the risk or benefit is not very large, even if the evidence would be convincing, or the opposite, that a risk or benefit may be large, even if the level of evidence is low. Findings like this can indicate crucial data gaps (Section 2.4) and may, in an objective way, help identify where further research is needed.

2.6. Substitution

In general, an RBA compares the health effects of two or more intake scenarios, defined as specified changes in the amount or type of food consumed. As a side effect, these specified changes in intake may also imply a change in the intake of other food products to compensate for the part of the diet that is deleted or added. So far, however, such “substitution” is rarely included in an RBA. The risks and benefits of increasing fish intake are for example frequently studied, but the related decrease in the intake of one or more other foods and the consequential health effects of that decrease are not included in the assessment (Berijka et al., 2012; Hoeckstra, Hart et al., 2013). Ideally, the risks and benefits of the change in intake in these other foods are included in the comparison of intake scenarios, but this severely complicates the RBA because it extends the list of risks and benefits to be included in the assessment. A complicating factor in this context is also that this substitution in terms of alternative amounts and types of foods eaten may vary among individuals, adding even more to the complexity of the RBA.

Alternatively, it can be that substitution is the specific purpose of the RBA, as for example in the case of food fortification, when a non-fortified food is replaced by a fortified food, and substitution is an inevitable part of the scenarios investigated (Hoeckstra et al., 2008). Likewise, substitution has been investigated in an RBA when added sugar is substituted by artificial sweeteners (Hendriksen, Tijhuis, Fransen, Verhagen, & Hoeckstra, 2011; Husey et al., 2008; Verhagen et al., 2012b). In the first case, no additional precautions need to be taken, as the fortified and non-fortified diets are similar except for the content of the specific nutrient. In the sugar-artificial sweetener case, the substitution leads to non-isocaloric diets and this may need to be addressed because it implies that the diet may change in more aspects than just the intended substitution.

To meet this challenge, it is a prerequisite that substitution is acknowledged in the RBA, either by specifically addressing it in the intake scenarios that are analysed, or by referring to it in the discussion of the assumptions and in the uncertainty characterization. As simplified substitution scenarios, one can consider replacements in the same food groups (e.g. meat and fish) and isocaloric alternatives (to make sure the energy intake stays similar). Next, the impact of substitution can be analysed in separate scenarios, where different options for substitution are compared.

2.7. The use of quantitative metrics

Within the tiered approach for RBA (Fransen et al., 2010; Hoeckstra et al., 2012), a qualitative approach can be sufficient if it is clear that the risks dominate the benefits or vice versa. If, alternatively, a quantitative approach is applied, the use of one common integrated health metric is needed to combine different positive and negative health
effects in an RBA and to compare different health effects within and between assessments. The quantitative metric that is used most in published RBAs of foods is the disability adjusted life years (DALY). The DALY is a measure that indicates how many healthy years of life are lost due to premature death or due to decreased quality of life associated with a disease or hazard (Devleeschauwer et al., 2014; Havelaar et al., 2000; Hoekstra et al., 2008; Murray, 1994). The quality of life is determined by the duration of illness and a weighing factor that indicates the severity of the specific disease considered (Salomon et al., 2015). The DALY is increasingly used for risk ranking (Van der Fels-Klerx et al., 2018) and in burden of disease studies (Havelaar et al., 2015), which aim to compare and prioritise health risks, it is used as an aid to policy makers when they have to decide where to spend their available resources. Methods used and results obtained in these studies are also useful for RBAs because the health effects considered can be the same and a large part of the underlying calculations is similar.

The DALY is commonly applied at a population level. Burden of disease, for example, is defined as the sum of individual DALY across the population, and applied as a measurement of the gap between current health status and an ideal health situation where the entire population lives to an advanced age, free of disease and disability (WHO, 2013). As risk-benefit questions are usually targeted at a change of intake scenario within the population (Section 2.3), the DALY is also commonly applied as a population metric in an RBA. However, populations consist of a large variety of individuals with varying food preparation habits, consumption patterns and sensitivity to food hazards. When the RBA is done and the risk-benefit balance for the population is interpreted as the risk-benefit balance for the average consumer, this does not mean that this balance is the same for all individual consumers. It can be that the balance goes in different directions for different subpopulations, e.g., the elderly, pregnant women or children, and because there are differences in intake and sensitivity between individuals. Therefore, the variability between consumers has to be taken into consideration, for example by using a stochastic approach (Hart et al., 2013).

Apart from the DALY, other metrics can be used, such as monetary integrated metrics like the cost-of-illness, which aims to calculate the direct and indirect monetary costs to society related to disease and death, or willingness-to-pay, a stated preference method which elicits the resources an individual is willing to give up for a reduction in a specific health risk. We refer to Mangen, Plass, & Kretzschmar, 2014, p. 196, for a comprehensive overview of these different metrics.

Even though the use of the DALY seems to be an established choice in RBAs, one should consider alternatives and remain critical on the choice of the preferred metric. Because this choice guides part of the data needs of the RBA and may have an impact on the interpretation of the final result, this choice should be made when the risk-benefit question is defined. As different metrics may convey different messages, the use of more than one metric could be considered as well. When metrics are used beyond the level of the general population, it is important to consider the impact of variability between consumers. Both the risk-benefit assessors and the decision makers should be aware of the strengths and weaknesses of the health metric chosen, as well as the underlying ethical dimensions (Arnesen & Kapiriri, 2004; Arnesen & Nord, 1999; Van der Fels-Klerx et al., 2018).

2.8. Including microbiology

As RBAs have predominantly been developed within the research areas of nutrition and toxicology, the concepts and definitions used are largely based on these two research areas (Section 1.1) and microbiology is not often included (Magnusson et al., 2012). Even though one of the first RBA publications relates to the risks and benefits of drinking water disinfection (Havelaar et al., 2000), only 7 of the 70 references indicated in the RBA review of Boué et al. (2015) include microbiology. Among those, there is only one from the BRAFO project, which, among topics not related to microbiology, discusses heat treatment of milk (Schütt et al., 2012). Microbiological benefits, e.g., the use of probiotic bacteria, have to our knowledge not yet been included in an RBA.

Reasons for this underrepresentation of microbiology in RBA are probably the intrinsic differences in the underlying research disciplines and the different nature of the associated health effects. Microbiological risks are often linked with mild health effects such as short episodes of gastro enteritis. They can also lead to long-term sequelae and severe chronic effects, but these are typically not registered and less often measured (Havelaar et al., 2012). In principle, microorganisms can rather easily be eliminated from foods by application of a heating process, which might suggest that microbiological risks from food can quite easily be prevented. However, microbial contamination of food products and exposure are common, and, to some extent, more easily accepted by consumers (Kher et al., 2013).

Burden of disease studies (Section 2.7) show an opposite trend compared with published RBA studies: because the availability of the relevant data is larger, the recent World Health Organization (WHO) study on the global burden of foodborne disease (Havelaar et al., 2015) is primarily focused on microbiological hazards, and only four chemical substances have been considered in the WHO report. The results suggest that the disease burden related to the exposure to microorganisms may be larger than that for chemicals, but more comparable disease burden estimates for chemical substances are required before an overall comparison between the burden of chemical substances and microbiological pathogens can be made. However, the results confirm that risk associated with microbiological hazards can be quantified and that it is important to include microbiological risks in RBAs as well.

The inclusion of microbiological risks and benefits in RBAs requires that the specific characteristics of microbiological agents are acknowledged, and that they are included in case studies. As illustrated by Berjia et al. (2012) microbiological risks can specifically be of importance when the effects of food processing are included in the risk-benefit question, as the doses largely depend on the storage and food preparation. It would therefore be advisable that data on food preparation (such as storage times, temperatures and the applied cooking style) are included in dietary surveys.

The challenges from differences in approach between chemical and microbiological risk assessment needs further study to allow the development of a more integrated approach towards RBAs (Sections 1.2 and 2.5). Recently developed tools that are increasingly adapted to allow comparisons between chemical and microbiological health risks (e.g. FDA-iRisk; Chen et al., 2013) can help to address these challenges.

2.9. The scope of risk-benefit assessments

The scope of a risk-benefit question in relation to food may be much wider than direct health impact and can include socio-economic, psychological and/or environmental dimensions (Boobis et al., 2013). When consumers select their food, the health effect is only one of the concerns; others include cost, taste, quality and sustainability of the production. An indicated health risk may be counterbalanced by each of these, for example, if low price and good taste are considered benefits that outbalance the health risk.

One may consider widening the scope of RBAs of foods and include some of the aspects mentioned above. Cost is an obvious choice, which is an intrinsic part of the RBA when metrics such as the cost-of-illness or willingness-to-pay are used (Section 2.7). It can also be added to the RBA by means of a cost-utility, cost-benefit or cost-effectiveness analysis, as for example done for the costs of intervention strategies that aim to lower the public health risks of Campylobacter from broiler meat (Mangen et al., 2007; Van Wagenberg, Van Horne, Sommer, & Nauta, 2016). Measurements such as the “cost per avoided DALY” can be highly informative for risk-benefit managers because they can indicate the economic consequences of scenarios in RBAs and allow for a comparison of policies.
Table 1
A summary of the challenges in risk-benefit assessment as discussed in this paper, with a brief indication of the proposed way forward.

<table>
<thead>
<tr>
<th>Topic</th>
<th>Challenge</th>
<th>Suggested way forward</th>
</tr>
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<tbody>
<tr>
<td></td>
<td>Definitions of basic concepts differ between disciplines underlying RBA. Risk and benefit assessments can be based on top-down human observational evidence or bottom-up risk assessment approaches, which may provide different health effect estimates of food compounds or contaminants.</td>
<td>Create awareness and reach consensus. Perform studies that combine the two approaches to compare potential bias and uncertainties, either by case studies or simulation studies.</td>
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<td></td>
<td>Risk-benefit question</td>
<td>Define the risk-benefit question in close collaboration with risk-benefit managers. Categorise questions and frame the risk-benefit question schematically.</td>
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<tr>
<td></td>
<td>Lack of data and knowledge; uncertainty</td>
<td>Identify, characterise and communicate uncertainties; fill up the crucial identified data gaps.</td>
</tr>
<tr>
<td></td>
<td>Imbalance of level evidence</td>
<td>Put emphasis on the size of the health effect rather than on the presence or absence of the health effect.</td>
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<tr>
<td></td>
<td>Substitution</td>
<td>Find a comparable food product and include it in the analysis, use isocaloric alternatives, or compare several scenarios.</td>
</tr>
<tr>
<td></td>
<td>Quantitative metrics</td>
<td>More than one metric can be useful, quantitative assessments can be preferable even if the risk-benefit balance is clear. Well balanced choices for the metrics applied have to be made when the risk-benefit question is defined.</td>
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<tr>
<td></td>
<td>Including microbiology</td>
<td>Perform more RBAs that include microbiological hazards, take advantage of experience in disease burden estimation and risk ranking.</td>
</tr>
<tr>
<td></td>
<td>Scope</td>
<td>Develop methods and metrics to do this further, integrate methods such as LCA and MCDA into RBAs.</td>
</tr>
<tr>
<td></td>
<td>Application</td>
<td>With case studies, show how useful the RBA can be in different areas and discuss experiences.</td>
</tr>
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</table>

Also, environmental sustainability can be taken into account, for example by the use of life cycle assessment (LCA), a product-oriented environmental assessment tool that provides a systematic way to quantify the environmental effects of individual products or services (Hermansen & Nguyen, 2012). A methodology is being developed to include nutritional health impacts in LCA (Stylianou et al., 2016), which could clearly contribute to the development of RBAs with a scope beyond immediate health effects of food intakes.

Ultimately, it can be attractive to address all of the relevant aspects in one overall analysis, for example by the use of multi criteria decision analysis (MCDA). This method has for example been applied to the prioritisation of foodborne pathogens (Ruzante et al., 2010), taking into account public health impact (expressed in DALY and cost-of-illness), market impact, consumer perception and acceptance, and social sensitivity to impacts on vulnerable consumer groups and industries. In MCDA, an integrating scoring method is developed, which weighs the importance of different factors that are considered relevant for the decision making, allowing one to come with a final ranking that includes all of these factors.

Defining the scope of the RBA is clearly an issue that should be decided upon when the risk-benefit question is formulated. A broader scope includes more relevant issues, but also implies an increasing demand for resources in terms of research efforts, data and method development. Clearly, challenges that complicate RBAs, such as the lack of data and knowledge, and the consequential uncertainties, the imbalance in level of scientific evidence and the use of quantitative metrics, only get more weight when a broader scope is taken. Yet, the ongoing developments show that progress can be made, and with multidisciplinary scientific collaboration and investment in research supporting RBAs, this progress can be strengthened in the future.

2.10. The application of risk-benefit assessments

So far, several RBAs have been performed, but mainly within research projects that were directed at the development of RBA frameworks and methodology. The aim of these RBAs was primarily to illustrate the potential of the methodology and the risk-benefit question was not posed by independent risk-benefit managers but by the researchers themselves. There is now a need for more experience with the practical application of RBAs and the proposed methodologies. These practical applications of RBAs can fall into two categories: those leading to recommendations or guidelines to food safety and health authorities, and those leading to process and formulation design by industry (Boué et al., 2015). The first application is the one considered most often and typically the request for such an RBA originates from national or international food and health authorities that have a mandate to advise the public on a particular food or diet and have identified a need to establish a scientific basis for this advice. Examples are an RBA on fish and fish products performed in Norway (Skåre et al., 2015) and an RBA on nuts performed in Denmark (Mejbø et al., 2015). Another reason for the authorities to make requests for an RBA is a need for an evaluation of health effects of proposed fortification of foods, as for example with vitamin D, folic acid (Hoekstra et al., 2008) or iodine (Zimmermann, 2008).

Food producers may have an interest in RBAs when they change their production or the formulation of their products. This is especially of interest when this change is based on a wish to decrease one specific health risk that can go at the expense of another. For example, when a heating step is introduced to decrease microbiological health risks, this can go at the expense of the formation of potentially carcinogenic substances (Havelaar et al., 2000) and/or decreased vitamin levels. In such cases, RBA can be an excellent tool to settle a dispute that cannot be solved on the basis of the identification of risks and benefits alone.

The challenge from increased application of RBAs can only be met by initiating more specific RBA projects based on current demands of risk-benefit managers and by performing RBAs in practice. Food safety and health authorities and the food industry should be open for multidisciplinary collaboration and should be made aware of the potential of RBAs. When RBAs are performed, they should be published in the international peer-reviewed literature, even if a lack of data or major uncertainties obstruct firm conclusions. This is important to assure the scientific quality, to increase the experience in the research community and to aid the international discussion on the potential and challenges of RBAs.
3. Conclusion

RBA is an evolving discipline in food safety and nutrition that takes advantage of achievements in a variety of underlying disciplines. As it integrates various health concerns, it is a valuable method to estimate the overall health effects related to food consumption and diet choice, which can be applied both by food and health authorities and the food industry. Recognizing the progress that has been made in the past decade and based on previous work, we have identified a series of challenges that should be met to develop the area further and indicated steps that should be taken for further progress. The challenges and suggested ways forward in meeting them are summarized in Table 1.

To meet the challenges of RBA, it is important that researchers in underlying disciplines and stakeholders in food regulation, production, retail and consumption from different regions in the world agree on definitions and concepts that are practical and agreeable for all. Based on relevant risk-benefit questions, a series of risk-benefit studies should be performed, not so much to develop methods, but predominantly to identify the practical challenges that are met when working on RBA case studies. When investigating these practical challenges, steps can be made in categorizing them and in developing and harmonizing agreeable methods to address them.

For the future development of the RBA area, it is important to perform methodological research into some of the identified challenges because they cannot be met by performing case studies alone. Examples are studies into the differences and similarities in results obtained from top-down compared with bottom-up approaches (by the application of comparative analytical tools and simulation studies), research into uncertainty analysis and comparative studies on integrated health metrics and metrics outside the health domain. Additionally, risk communication is one of the key pillars in risk analysis and should also be an inherent part of RBAs of foods, particularly for the communication of quantitative metrics and their attending uncertainties to all stakeholders.

Overall, with an increasing demand from different stakeholders for holistic and objective assessments of the health effect of foods, multi-disciplinary RBA is a promising research area for the future. Impressive progress has been made and, despite the remaining challenges, we expect that more progress will be made in the next decade. The steps forward proposed in this paper will be useful in taking the research area further, allowing for transparent and reliable RBAs to be performed on a wider scale in the future.

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References
